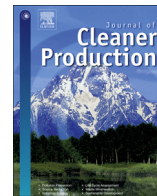




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Stormwater wetlands for the enhancement of environmental ecosystem services: case studies for two retrofit wetlands in Brisbane, Australia

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ABSTRACT

Constructed stormwater wetlands are examples of 'blue-green' ecological engineering in the urban environment providing a range of ecosystem services. A study was undertaken in Brisbane, Australia to examine the enhancement of environmental ecosystem services in 2 retrofit wetlands. Originally concrete drainage channels in existing residential catchments, a series of wetlands and ponds were created i.e. green infrastructure in a brownfield setting. Boardwalks, pathways and interpretive signage were incorporated into the design.

Golden Pond wetland was only 0.2% of the catchment whereas Bridgewater Creek – Bowie's Flat wetland was 0.4% of the catchment. The larger 0.8 ha Bridgewater Creek wetland had an overflow by-pass channel. Water quality (suspended solids and nutrients) and the diversity of aquatic macro-invertebrates were monitored for 2–4 years after construction.

During storm events suspended solids increased at Golden Pond due to resuspension from the high velocities in this small linear flow system. Resuspension was minimised at Bridgewater Creek due to the by-pass channel. During dry weather suspended solids in both wetlands increased due to resuspension from the activity of waterbirds. Both wetlands were effective in reducing nutrient concentrations especially $\text{NO}_x\text{-N}$ and $\text{PO}_4\text{-P}$, attributed to biological uptake; but during dry periods $\text{NH}_4\text{-N}$ increased within the system. The concept of irreducible background concentrations in wetland systems is discussed.

Open water and aquatic vegetation provided habitats for macroinvertebrates. Species richness increased by more than 50% in both stormwater wetlands, compared to the concrete drains.

Wetland sizing and design are essential parameters for water quality improvement and biodiversity.

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1. Introduction

Urbanisation decreases both green (forests and fields) and blue (wetlands and water bodies) space and increases impervious surface area so that rainfall infiltration capacity is reduced and the volume of stormwater runoff is increased. This leads to higher flood peaks in urban waterways with possible flooding and higher loads (concentrations and mass) of contaminants: sediment, nutrients, metals, hydrocarbons, organic micro-pollutants i.e. xenobiotics, pathogens. Many of these contaminants are primary stressors of aquatic ecosystem health (Greenway, 2004; Kadlec and Wallace, 2009; Hvitved-Jacobsen et al., 2010).

Whilst the potential impacts of stormwater runoff, as a non-point source of pollution on water quality (and potential ecosystem health) have been recognised since the 1980's, legislation on stormwater control and the implementation of best management practices (BMP's) were not widely recognised and accepted until the 1990's (US EPA, 1996, 2000; Roy et al., 2008; Hvitved-Jacobsen et al., 2010). In USA best management practises are referred to as low impact development (LID), in Australia as water sensitive urban design (WSUD) and in UK as sustainable urban drainage systems (SUDS).

In Australia the first guidelines for Water Sensitive Urban Design (WSUD) were in Western Australia (Whelans et al., 1994). In Queensland, Brisbane City Council released similar guidelines in 1999 (BCC, 1999a,b). Brisbane City Council (2000) set stringent Water Quality Objectives: 15 mg L^{-1} TSS, 5 mg L^{-1} TVS, 0.65 mg L^{-1} TN, 0.035 mg L^{-1} $\text{NH}_4\text{-N}$, 0.13 mg L^{-1} $\text{NO}_3\text{-N}$, 0.07 mg L^{-1} TP,

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0.035 mg L⁻¹ PO₄-P, 8 µg L⁻¹ chlorophyll-a; to meet legislative requirements to improve the quality of stormwater runoff which ultimately flows into Moreton Bay, designated in 1993 as a RAMSAR wetland/marine park with significant fisheries values and international ecological values ([Ramsar Information Sheet: Moreton Bay](#)).

Structural BMPs include: Gross Pollutant Traps (to catch coarse sediment and trash), Retention/Detention/Sediment Basins (to capture coarse and fine sediment, and removal of soluble contaminants), Vegetation Buffer Strips (removal by sheet flow across wide natural vegetation strips), Infiltration and Bioretention Systems (removal by physical filtration, chemical and biological processes), Vegetation Filter Strips/Grass Swales (removal along concentrated flow paths), Water Quality Control Ponds/Wet Ponds/Dry Ponds/Wetlands (aquatic ecosystems – contaminant removal assisted by wetland plants, algae and microorganisms) ([BCC, 1999a,b](#); [US EPA, 2000](#); [Hvitved-Jacobsen et al., 2010](#)). Although the primary purpose of BMP's was seen as hydrological-controlling runoff flows, flood mitigation, water storage and contaminant removal to improve downstream water quality, it is now recognised that BMP's can be an infrastructure opportunity for the improvement of environmental (habitat and landscape enhancement, biodiversity, ground water protection) and social (recreational amenities, aesthetics, sustainable living) services in urbanised areas ([Martin et al., 2007](#); [Roy et al., 2008](#); [Marsalek and Schreier, 2009](#); [Hvitved-Jacobsen et al., 2010](#); [Barbarosa et al., 2012](#)).

Wetlands and ponds are suitable BMP's for both water storage and water quality improvement. Vegetation (emergent aquatic plants) is the dominant feature of wetlands, whereas open water is the dominant feature of ponds ([Greenway et al., 2007](#)). Ponds, which include sediment basins, retention/detention ponds are often deeper than 1.5 m and emergent plants are restricted to the shallow littoral margins. Submerged pondweeds may occur if there is sufficient light. The design of stormwater wetlands usually incorporates both zones of dense vegetation (shallow macrophytes zones) and deeper open water ([Greenway, 2004](#); [Kadlec and Wallace, 2009](#)). The 'treatment' of stormwater as it flows through a wetland/pond is the result of complex interactions between physical, chemical and biological processes ([Greenway, 2004](#); [Kadlec and Wallace, 2009](#); [Hvitved-Jacobsen et al., 2010](#)). Vegetation plays an important role in these processes.

Ponds and wetlands also provide habitats for a range of biota, aquatic plants and animals as well as terrestrial wildlife ([Wetzel, 2001](#); [Mitsch and Gosselink, 2007](#); [Cereghino et al., 2014](#)). When constructed in residential areas to treat stormwater these water bodies can potentially balance the detrimental effects of urbanisation on predevelopment landscapes and ecosystems, achieving and sustaining biodiversity. Only recently have studies assessed these ecological values in stormwater wetlands ([Woodcock et al., 2010](#); [Tixier et al., 2011](#); [Moore and Hunt, 2012](#), [Herrmann, 2012](#)).

This paper examines the environmental services of water quality improvement and aquatic macroinvertebrate biodiversity provided by 2 stormwater wetlands constructed within existing older residential areas ie new 'green and blue' infrastructure within old 'brownfield' settings. Both wetlands were created from retrofitting concrete drainage channels. Social values are briefly discussed.

2. Materials and methods

2.1. Wetland descriptions

Two urban stormwater wetland systems were selected with similar catchment size (180 ha) to monitor water quality treatment and aquatic macroinvertebrate biodiversity. Both systems are located in Brisbane, SE Queensland; Australia. The climate is subtropical with an average annual rainfall of 1120 mm. The wet season

occurs over summer/autumn, with an average of 10 days with greater than 50 mm daily rainfall; intense rainfall events of 1 h duration of 71 mm/h are common. During the study period a 1 in 100 year event occurred when 208 mm fell in 3 h.

The significance of the wetlands selected is that they are both 'retrofit' systems. They were designed and constructed within existing residential estates (brownfield sites) using sections of concrete channelised streambeds.

2.1.1. Golden Pond stormwater treatment wetland system

The Golden Pond system is located in Calamvale—a new (1990's) outer suburb SW of Brisbane. Prior to development three stream channels (tributaries of Calamvale Creek) fed into a series of lagoons with fringing riparian wetlands – known as the Kameruka Wetlands. The main north-south flowing stream tributary has a catchment of 160 ha, whereas the west-east flowing stream tributaries have a combined catchment of 50 ha.

As a consequence of residential development, considerable modification occurred. Initially the main (north-south flowing) stream was channelised into a concrete lined trapezoidal channel. However in 1999 Brisbane City Council decided to 'retrofit' the lower section (120 m) of the concrete channel into a 'linear flow wetland system' including a small sediment basin (20 m × 13 m), a shallow wetland (Wetland1 80 m × 20 m) and a deeper pond (Wetland2 50 m × 20 m) ([Fig. 1](#)). Pathways and a picnic shelter in parkland adjacent to Wetland 2 were incorporated into the design. Interpretive signage was erected to increase public awareness of stormwater quality improvement devices (SQIDs).

Stormwater from the concrete sediment basin flows into Wetland 1 (the Wetland) via a v-notched weir. Base flow and storm event discharge rates are given in [Table 1](#). Discharge rates calculated for stormwater leaving the sediment basin and entering Wetland 1 ranged from 3 to 5.7 m³ s⁻¹ for extreme (>20 y ARI) storm events, and from 0.15 to 0.8 m³ s⁻¹ for high-intensity rain squalls. At discharge rates greater than 0.45 m³ s⁻¹, short-circuiting occurs through the middle due to the positioning of a single V-notch weir, the lack of dense emergent macrophytes, and the linear nature of the flow path through the wetland. There is no bypass channel.

Between Wetland 1 (the Wetland) and Wetland 2 (the Pond), the water flows over a wide concrete sill. Piped underground drainage from the western section of residential development (20 ha) passes through a below ground Gross Pollutant Trap (GPT): CDS Proprietary Unit, before being discharged into Wetland2. Water then flows into Calamvale Creek.

The narrow outlet (1 m width) at the end of Wetland 2 ensures that the water backs up, thereby increasing the retention time. Water storage capacity and detention times are given in [Table 1](#). It has been estimated that the average retention time for both wetlands during non-extreme storm events would be between 3 and 5 h, and between 5 and 32 h for less intense rainfall events.

The total area of this 'constructed wetland system' is 3000 m² i.e. 0.2% of the watershed catchment.

Piped drainage from the south western section of residential development (30 ha) also passes through a below ground GPT: ECOSOL Proprietary Unit, before entering the upper reaches of Calamvale Creek within a densely vegetated riparian wetland.

Calamvale Creek, several lagoons and adjacent riparian vegetation - the Kameruka Wetlands were retained in their natural state for a length of 680 m ([Fig. 1](#)). Calamvale Creek flows into Scrubby Creek, then the Logan River and into Moreton Bay Ramsar Wetland.

2.1.2. Bridgewater Creek stormwater treatment wetland system: Bowie's Flat Wetland

The Bridgewater Creek system is located at Cooparoo (SE Brisbane). Urbanisation of this suburb and adjacent Camp Hill occurred

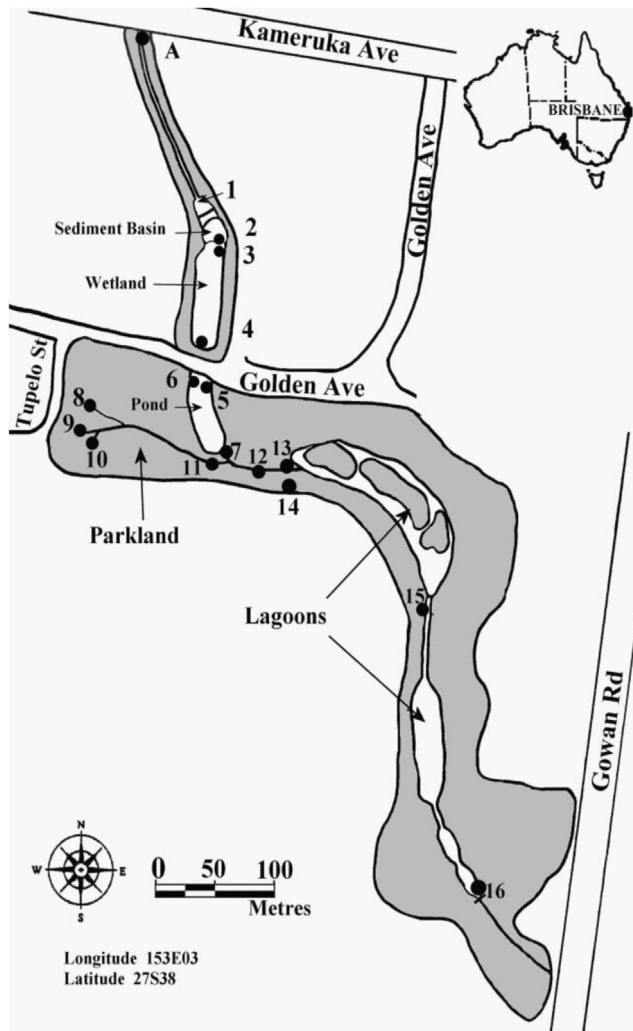


Fig. 1. Location Map of Sampling Sites (1–16) at Golden Pond Stormwater Treatment System (sediment basin; wetland = Wetland1; pond = Wetland2) and Kameruka Wetlands (Lagoons), Calamvale, Brisbane.

prior to 1950, when the original Bridgewater Creek was realigned and channelised to form a concrete drain. Bridgewater Creek flows into Norman Creek, then into the Brisbane River and into Moreton Bay Ramsar Wetland.

The Bridgewater Creek stormwater treatment wetland was constructed in Bowie's Flat parkland (an area of 1.5 ha). The

concrete drain and adjoining grassy open space were excavated for the wetland basin (0.8 ha). While water quality improvement was the primary objective of this wetland, a picnic shelter, boardwalks and pathways for enhanced community visitations and recreation were incorporated into the design (*Water by Design-Bridgewater Creek Wetland*). Interpretive signage was erected to increase public awareness of the value of wetlands for water quality improvement and to enhance biodiversity.

Construction was completed in October 2001. This wetland is locally known as Bowie's Flat Wetland.

The "wetland" system consists of six inter-connected ponds and receives stormwater from two tributaries (Fig. 2-inlets). The main south-north flowing stream channel (Bridgewater Creek) has a catchment of 140 ha, whereas the piped drainage system from the west collects runoff from a 40 ha catchment. Before entering the "wetland", the stormwater runoff from both tributaries passes through trash racks, i.e. above-ground GPT, to collect trash and organic debris (leaf litter). The stormwater then passes into a sediment basin (Pond 1).

Pond 1, the sediment basin, is triangular in shape, with a surface area of 1000 m², a depth of 2 m and a volume of 2000 m³ (standing water level) with a 2500 m³ capacity (during storm events). It has a fringing littoral zone of sedges.

Ponds 2 to 6 are oval-shaped, with dimensions of 40–50 m in length and 15–20 m in width; 0.2–2 m depth. They were originally designed as "macrophyte zones" to include open water, deep marsh, shallow marsh and ephemeral zones. Unfortunately, the establishment of wetland vegetation has been poor and is now largely restricted to narrow fringes along the shallow edges of the ponds (Greenway et al., 2007; Hunt et al., 2011).

Ponds 2–6 have a combined surface area of 7000 m², and a volume of 4000 m³ (standing water level) and 10000 m³ (during storm events). The total area of this 'constructed wetland system' ('ponds' 1–6) is 8000 m²–0.4% of the watershed catchment.

Water flow from Pond 1 to Pond 2 occurs via an underground pipe, whereas surface water flow occurs progressively between Pond 2 to Pond 6. During periods of light rainfall and during dry weather, stormwater flows progressively through Ponds 1 to 6. In times of heavy rainfall or storm events, the water levels rise rapidly so that Ponds 2 to 6 form a single lake-like water body. The outlet riser structure in Pond 6 determines stormwater residence time in the wetland and has been designed to ensure stormwater is retained in the wetland for at least 48 h optimum detention time for treatment determined by Brisbane City Council (BCC, 1999a,b). The retention times during wet weather range from 36 h for a major storm event to 6 days for less intense rainfall (Table 1). However at maximum ponding water level Bayley (2007) calculated a 9day extended detention time.

Table 1

Comparison of size, water storage, discharge rates (from sediment basin/Pond1) and detention times of the Golden Pond and Bridgewater Creek stormwater wetland systems.

	Golden pond (0.2% catchment)		Bridgewater Creek (0.44% catchment)	
	Sediment basin	Wetland 1 and 2 'wetland' & 'pond'	Pond 1 sediment basin	Ponds 2–6 'wetlands'
Vegetation	Absent	water lilies, aquatic creepers, submerged pond weeds	littoral/marsh emergent sedges submerged pond weeds in pond 6	
Area	284 m ²	2600 m ²	1000 m ²	7000 m ²
Depth	1.5 m	0.2–1.2 m	2 m	0.2–1.5 m
Volume: Standing Water Level	100 m ³	1300 m ³	2000 m ³	4000 m ³
Storm event	150 m ³	2000 m ³	2500 m ³	10,000 m ³
Discharge rates:				
Storm events	4.22–0.15 m ³ s ⁻¹		2.33–0.33 m ³ s ⁻¹	
Base flow	0.0015–0.0003 m ³ s ⁻¹		0.0015–0.0003 m ³ s ⁻¹	
Detention:	3–16 h storm event;		36 h – 6 days storm event;	
Base flow	16–20 days dry base flow		>20 days dry base flow	

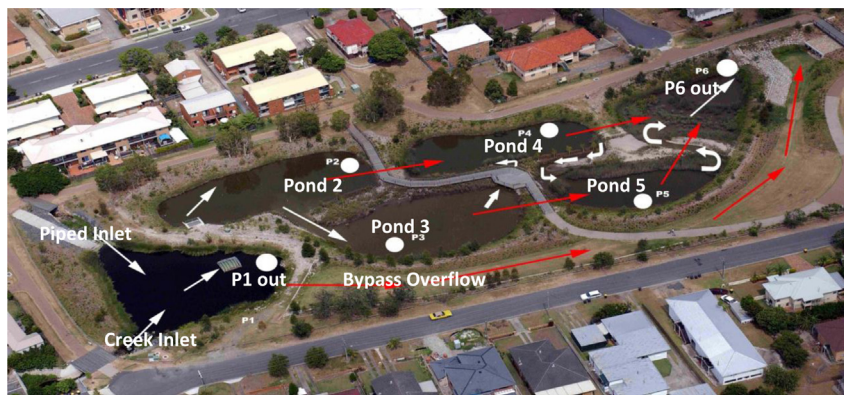


Fig. 2. Aerial view of the Bridgewater Creek constructed wetland system: Bowie's Flat Wetland showing sampling sites (P1out:Pond1-sediment basin; P2: Pond2; P3:Pond3; P4:Pond4; P5:Pond5; P6out:Pond6). The white arrows depict base flow and stormwater flow during 'average rainfall' events. The red arrows depict stormwater flow during 'high intensity rainfall' events. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

During high-intensity storm events with high flow, the stormwater from Pond 1 can overflow via a spillway into a grassed "overflow bypass channel" which flows around the "wetland" and re-enters Bridgewater Creek flow path downstream of Pond 6 outlet. Bayley (2007) calculated that an intense 10 mm rainfall event in the catchment would generate runoff volumes greater than storage capacity and trigger the overflow from Pond1. The bypass is designed to protect the wetland vegetation from being damaged during high flow events and to reduce resuspension of the sediment. Since the overflow pathway is a grass swale it provides some sediment retention.

2.2. Water quality monitoring

Urbanised catchments with large impervious areas respond quickly to rainfall events. This often makes it difficult to interpret/correlate water quality data with rainfall and runoff, unless synchronisation of sampling occurs. Due to the large number of sampling sites in these 2 systems automated samplers could not be installed, instead grab samples were taken. However, during storm events this presented difficulties since all sites could not be sampled simultaneously and many storms occurred at night. Thus 'storm event' samples were classified as "wet weather", with samples being collected within 12 h of the storm. Inter-event- base flow samples were classified as "dry weather". Data presented in this paper are from water quality monitoring over a 2 year period from 2002 to 2004.

At Golden Pond 6 Wet and 14 Dry weather water samples were collected. At Bridgewater Creek 20 Wet and 26 Dry weather samples were collected.

All water samples were analysed for Total Suspended Solids (TSS) and Total Volatile Solids (TVS) – the organic fraction; and nutrients: ammonium $\text{NH}_4\text{-N}$; nitrate-nitrite $\text{NO}_x\text{-N}$; total nitrogen TN; phosphate $\text{PO}_4\text{-P}$ and total phosphorus TP. The analyses were carried out according to the standard methods (APHA-AWWA-WEF, 2005). TSS measurements followed procedure number 2540 D (gravimetric, dried at 105 °C) and TVS tests followed procedure number 2540 E (gravimetric, ignited at 550 °C). Nutrients were analysed using colorimetric methods with a Lachat Quikchem 8000 Flow Injection Analyser (FIA) (Hach Company, Loveland, Colorado). Subsamples for soluble nutrients were filtered with a 0.45-mm filter paper before refrigeration. Total N and P were analysed following standard persulfate digestion on unfiltered samples.

Chlorophyll-a was measured (APHA-AWWA-WEF, 2005) as an indicator of phytoplankton activity.

2.3. Macroinvertebrate monitoring

Macroinvertebrates were sampled using a D-frame sweep net (0.5 mm mesh) which was alternately 'jabbed' into clumps of wetland vegetation and 'swept' through the water column over a distance of 1–2 m. Monitoring occurred over 2 years at Golden Pond and 4 years at Bridgewater Creek. Both summer (wet season) and winter (dry season) sampling was conducted. A 200 mL scoop was used to monitor the presence and abundance of any mosquito larvae (Greenway et al., 2003).

3. Results

3.1. Water quality

3.1.1. Total suspended solids and Total Volatile Solids

Tables 2 and 3 present both Wet and Dry weather data for the Golden Pond and Bridgewater Creek systems respectively. At Golden Pond the small sediment basin was not effective in removing TSS and TVS. At the bottom of Wetland1 TSS and TVS were higher than the stormwater both entering and leaving the sediment basin ie export; the increase in Wetland1 suggests resuspension. In storm events discharge rates from the V-notch weir in the sediment basin were as high as $4.22 \text{ m}^3 \text{ s}^{-1}$ (Table 1). Wetland2 removed some of the TSS but outflow concentrations were still higher than the stormwater inflows. Thus the Golden Pond constructed wetland system was not effective in removal of TSS and TVS during storm events.

The natural riparian wetland removed TSS from the Ecosol in Wet, but TSS increased in Dry due to an increase in organics-TV. TSS in the Kameruka Wetlands showed large variation, but at 600 m downstream TSS decreased. In Dry TSS was lower than the inflow stormwater upstream. The proportion of TVS in Dry was consistently higher, and in the natural downstream channel made up to 75%. The natural Kameruka Wetlands were more beneficial in TSS removal than the constructed wetlands in both storm events and base flow.

TSS and TVS in stormwater entering the Bridgewater Creek system showed large variations and was higher than stormwater runoff in the Golden Pond catchment. TSS leaving the sediment basin (Pond1 outlet) was lower than the stormwater inflows in both Wet and Dry i.e. there was removal. TVS was higher (70%) in Dry possibly due to phytoplankton. In Wet TSS between Pond 1 and Pond 6 decreased with an overall removal of 27%. However, in Dry TSS and TVS increased in Ponds 2 and 3 due to particulate organics.

Table 2
TSS and TVS (mg/L mean \pm SD) Golden Pond Stormwater Treatment Train: ns (insignificant).

Site	Description	Wet TSS	Wet TVS	% Organics	Dry TSS	Dry TVS	% Organics
1	In Sediment Basin	17 \pm 12	4 \pm 3	41%	6 \pm 4	3 \pm 2	50%
2	Out Sediment Basin	20 \pm 8	5 \pm 2	25%	6 \pm 3	3 \pm 3	50%
	1–2: Removal	ns	ns		ns	ns	
3	Top Wetland 1	22 \pm 9	5 \pm 3	23%	7 \pm 4	4 \pm 2	57%
4	Bottom Wetland 1	26 \pm 10	6 \pm 3	23%	14 \pm 6	6 \pm 3	43%
	3–4: Removal	EXPORT	ns		EXPORT	EXPORT	
5	Top Wetland 2	53	11	22%	16 \pm 9	8 \pm 7	50%
6	CDS S/W Outlet	12 \pm 2	4 \pm 2	33%	6 \pm 3	4 \pm 2	67%
7	Bottom Wetland 2	24 \pm 12	5 \pm 3	21%	13 \pm 9	7 \pm 5	54%
	4–7: Removal %	8%	17%		7%	ns	
	1–7:	EXPORT	EXPORT		EXPORT	EXPORT	
9	Ecosol Bypass	23 \pm 9	5 \pm 1	22%	7 \pm 5	4 \pm 3	57%
10	Ecosol S/W Outlet	18 \pm 8	5 \pm 2	28%	12 \pm 9	5 \pm 4	42%
11	Riparian Wetland	15 \pm 8	4 \pm 2	27%	14 \pm 11	7 \pm 5	50%
	9–11: Removal %	35%	ns		EXPORT	EXPORT	
12	Confluence 7 + 11	24 \pm 16	5 \pm 4	21%	9 \pm 5	6 \pm 4	67%
13	100 m downstream	29 \pm 9	6 \pm 2	21%	6 \pm 4	4 \pm 2	67%
15	230 m downstream	37 \pm 7	10 \pm 4	33%	8 \pm 5	6 \pm 4	75%
16	600 m downstream	21 \pm 16	5 \pm 4	24%	5 \pm 2	3 \pm 3	60%
	12–16: Removal %	13%	nil		44%	50%	

Table 3
TSS and TVS (mg/L mean \pm SD) Bridgewater Creek Stormwater Treatment System.

Location	Wet weather			Dry weather		
	TSS	TVS	% TVS	TSS	TVS	% TVS
Creek Inlet	66 \pm 40	21 \pm 14	32%	18 \pm 17	5 \pm 3	28%
Piped Inlet	33 \pm 16	10 \pm 4	30%	17 \pm 16	5 \pm 3	30%
Pond 1 out	41 \pm 34	9 \pm 5	22%	10 \pm 6	7 \pm 5	70%
IN-OUT P1	Decrease			Decrease		
Pond 2	40 \pm 27	11 \pm 5	28%	30 \pm 20	18 \pm 12	60%
Pond 3	40 \pm 27	14 \pm 8	35%	37 \pm 20	27 \pm 17	73%
Pond 4	35 \pm 27	15 \pm 6	43%	23 \pm 10	16 \pm 7	70%
Pond 5	34 \pm 28	13 \pm 8	38%	20 \pm 10	12 \pm 7	60%
Pond 6 out	30 \pm 26	10 \pm 5	33%	16 \pm 8	7 \pm 4	44%
P1–P6%removal	27%			Export		
IN-OUT P1	Decrease			Decrease		

Although some reduction occurred in Ponds 4 and 5, outflow from Pond 6 was still higher than from Pond 1. However overall there was a slight decrease in TSS compared to the stormwater inflows. The proportion of TVS in Dry was consistently higher than in Wet suggesting either phytoplankton or resuspension of organic matter.

Table 4
Nutrients (mg/L mean \pm SD): Golden Pond Stormwater Treatment Train: Wet Weather.

Site	Description	NH ₄ -N	NO _x -N	TN	PO ₄ -P	TP
1	Into Sed Basin	0.10 \pm 0.10	0.83 \pm 0.48	1.27 \pm 0.63	0.09 \pm 0.09	0.11 \pm 0.05
2	Out Sed Basin	0.09 \pm 0.09	0.68 \pm 0.38	1.22 \pm 0.58	0.08 \pm 0.04	0.12 \pm 0.05
	1–2: Removal%	10%	18%	4%	11%	insignificant
3	Top Wetland 1	0.12 \pm 0.10	0.47 \pm 0.28	1.01 \pm 0.28	0.08 \pm 0.07	0.11 \pm 0.02
4	Bottom Wetland 1	0.08 \pm 0.07	0.43 \pm 0.46	1.12 \pm 0.58	0.06 \pm 0.06	0.13 \pm 0.07
	3–4: Removal%	33%	9%	export	25%	insignificant
	1–4: Removal%	20%	40%	15%	34%	
5	Top Wetland 2	0.09 \pm 0.06	0.25 \pm 0.40	0.92	0.03 \pm 0.04	0.09 \pm 0
6	CDS S/W Outlet	0.12 \pm 0.05	1.92 \pm 1.47	2.50 \pm 1.22	0.24 \pm 0.15	0.34 \pm 0.22
7	Bottom Wetland 2	0.07 \pm 0.04	0.37 \pm 0.28	0.94 \pm 0.51	0.05 \pm 0.05	0.11 \pm 0.06
	6–7: Removal%	42%	80%	62%	79%	65%
9	Ecosol Bypass	0.09 \pm 0.05	0.53 \pm 0.60	NA	0.07 \pm 0.06	0.07
10	Ecosol S/W Outlet	0.11 \pm 0.08	1.07 \pm 0.68	1.75 \pm 0.55	0.10 \pm 0.09	0.15 \pm 0.06
11	Riparian Wetland	0.05 \pm 0.04	0.61 \pm 0.45	1.14	0.07 \pm 0.08	0.13 \pm 0.04
	10–11: Removal%	55%	43%	35%	30%	13%
12	Confluence 7 + 11	0.07 \pm 0.03	0.68 \pm 0.35	1.16 \pm 0.46	0.06 \pm 0.08	0.13 \pm 0.04
13	100 m downstream	0.13 \pm 0.01	0.63 \pm 0.15	1.22 \pm 0.29	0.05 \pm 0.05	0.10 \pm 0.02
15	230 m downstream	0.12 \pm 0.03	0.47 \pm 0.16	0.71	0.03 \pm 0.04	0.12 \pm 0.02
16	600 m downstream	0.06 \pm 0.03	0.42 \pm 0.17	0.92 \pm 0.27	0.04 \pm 0.02	0.09 \pm 0.03
	12–16: Removal%	14%	38%	21%	33%	31%

3.1.2. Nutrients

Tables 4–7 present Wet and Dry weather nutrient data for the Golden Pond and Bridgewater Creek systems respectively. Both systems show large variations in stormwater inflow nutrient concentrations in Wet and Dry, and variation within the different 'wetland components' however, mean outflow concentrations were consistently lower than stormwater entering the systems.

At Golden Pond in Wet there was negligible removal of nutrients into and out of the sediment basin; however significant removal of NH₄-N, NO_x-N and PO₄-P occurred in Wetland 1. Wetland 2 was highly effective in removing nutrients flowing into the wetland via the piped drainage system (CDS Stormwater Outlet) with up to 80% reduction in NO_x-N and PO₄-P concentrations and 65% reduction in TN and TP concentrations. The riparian wetland was also effective in reducing nutrient concentrations via the Ecosol Stormwater Outlet. The 600 m length of natural wetlands was effective in further nutrient reduction.

In Dry only a reduction in PO₄-P concentrations occurred in the sediment basin; reductions of NH₄-N, NO_x-N and PO₄-P occurred in Wetland 1. Wetland 2 was again highly effective in removing nutrients from the S/W Outlet, though NH₄-N was higher at

Table 5
Nutrients (mg/L mean \pm SD): Golden Pond Stormwater Treatment Train: Dry Weather.

Site	Description	NH ₄ -N	NO _x -N	TN	PO ₄ -P	TP
1	Into Sed Basin	0.03 \pm 0.03	0.53 \pm 0.67	0.57 \pm 0.31	0.04 \pm 0.04	0.08 \pm 0.02
2	Out Sed Basin	0.05 \pm 0.05	0.56 \pm 0.71	0.70 \pm 0.33	0.03 \pm 0.03	0.08 \pm 0.01
3	Top Wetland 1	0.04 \pm 0.03	0.43 \pm 0.63	0.74 \pm 0.33	0.03 \pm 0.03	0.07 \pm 0.01
4	Bottom Wetland 1	0.03 \pm 0.03	0.25 \pm 0.49	0.63 \pm 0.33	0.02 \pm 0.02	0.07 \pm 0.02
5	Top wetland 2	0.06 \pm 0.09	0.31 \pm 0.41	0.70 \pm 0.28	0.02 \pm 0.03	0.09 \pm 0.02
6	CDS S/W Outlet	0.07 \pm 0.06	1.76 \pm 0.78	2.35 \pm 0.37	0.11 \pm 0.06	0.19 \pm 0.05
7	Bottom Wetland 2	0.08 \pm 0.09	0.25 \pm 0.23	0.97 \pm 0.50	0.05 \pm 0.08	0.14 \pm 0.07
9	Ecosol Bypass	0.02 \pm 0.02	0.30 \pm 0.51	1.05 \pm 0.45	0.01 \pm 0.01	0.09 \pm 0.04
10	Ecosol S/W Outlet	0.17 \pm 0.36	1.10 \pm 0.81	1.53 \pm 0.79	0.04 \pm 0.04	0.08 \pm 0.02
11	Riparian Wetland	0.26 \pm 0.43	0.28 \pm 0.25	0.84 \pm 0.25	0.02 \pm 0.02	0.09 \pm 0.05
12	Confluence 7 + 11	0.09 \pm 0.06	0.21 \pm 0.18	0.75 \pm 0.32	0.03 \pm 0.03	0.08 \pm 0.04
13	100 m downstream	0.06 \pm 0.06	0.22 \pm 0.28	0.59 \pm 0.30	0.02 \pm 0.02	0.06 \pm 0.02
15	230 m downstream	0.07 \pm 0.06	0.08 \pm 0.11	0.61 \pm 0.21	0.02 \pm 0.02	0.05 \pm 0.04
16	600 m downstream	0.03 \pm 0.03	0.09 \pm 0.11	0.60 \pm 0.28	0.02 \pm 0.02	0.05 \pm 0.03
	12–16: Removal%	67%	59%	20%	33%	38%

Table 6
Nutrients: Bridgewater Creek treatment system: Wet weather (mean \pm SD).

Site	NH ₄ -N (mg/L)	NO _x -N (mg/L)	TN (mg/L)	PO ₄ -P (mg/L)	TP (mg/L)
Creek Inlet	0.14 \pm 0.12	0.43 \pm 0.17	4.68 \pm 2.70	0.14 \pm 0.16	0.70 \pm 0.38
Piped Inlet	0.12 \pm 0.10	0.42 \pm 0.32	3.62 \pm 3.36	0.10 \pm 0.09	0.31 \pm 0.22
Pond 1 Out	0.13 \pm 0.11	0.46 \pm 0.37	2.69 \pm 3.12	0.15 \pm 0.21	0.21 \pm 0.19
IN – OUT P1	insignificant	insignificant	decrease	insignificant	decrease
Pond 2	0.31 \pm 0.03	0.13 \pm 0.07	1.14 \pm 0.19	0.04 \pm 0.02	0.20 \pm 0.03
Pond 3	0.10 \pm 0.10	0.13 \pm 0.14	1.84 \pm 0.87	0.08 \pm 0.05	0.30 \pm 0.14
Pond 4	0.20 \pm 0.27	0.16 \pm 0.10	1.73 \pm 0.37	0.04 \pm 0.03	0.26 \pm 0.18
Pond 5	0.28 \pm 0.06	0.16 \pm 0.10	1.42 \pm 0.31	0.02 \pm 0.02	0.19 \pm 0.06
Pond 6 Out	0.13 \pm 0.22	0.10 \pm 0.08	0.84 \pm 0.21	0.02 \pm 0.02	0.12 \pm 0.08
P1–P6: Removal%	insignificant	78%	69%	87%	43%

Table 7
Nutrients: Bridgewater Creek treatment system: Dry weather (mean \pm SD).

N	NH ₄ -N (mg/L)	NO _x -N (mg/L)	TN (mg/L)	PO ₄ -P (mg/L)	TP (mg/L)
Creek Inlet	0.08 \pm 0.09	0.57 \pm 0.60	1.84 \pm 1.01	0.19 \pm 0.18	0.26 \pm 0.11
Piped Inlet	0.06 \pm 0.05	1.10 \pm 0.40	1.97 \pm 0.28	0.16 \pm 0.10	0.24 \pm 0.09
Pond 1 Out	0.10 \pm 0.09	0.12 \pm 0.24	1.28 \pm 0.45	0.08 \pm 0.06	0.22 \pm 0.10
IN–OUT P1	increase	decrease	decrease	decrease	decrease
Pond 2	0.22 \pm 0.30	0.06 \pm 0.10	1.46 \pm 0.35	0.05 \pm 0.04	0.27 \pm 0.06
Pond 3	0.22 \pm 0.29	0.07 \pm 0.17	1.38 \pm 0.50	0.04 \pm 0.03	0.26 \pm 0.08
Pond 4	0.15 \pm 0.23	0.03 \pm 0.04	1.14 \pm 0.35	0.03 \pm 0.02	0.23 \pm 0.07
Pond 5	0.19 \pm 0.23	0.08 \pm 0.15	1.23 \pm 0.38	0.02 \pm 0.02	0.21 \pm 0.08
Pond 6 Out	0.11 \pm 0.11	0.10 \pm 0.14	1.04 \pm 0.36	0.02 \pm 0.01	0.17 \pm 0.07
P1–P6 out	Export	17%	19%	75%	23%
IN–OUT P6	EXPORT	82–91%	44–47%	88–90%	35–29%

Wetland2 outlet. Similarly the riparian wetland reduced nutrients from the S/W outlet but NH₄-N increased. The downstream wetland provided further nutrient reduction including NH₄-N (67%).

Thus both the constructed wetland system and the natural wetland system were effective in nutrient removal from stormwater inflows.

At Bridgewater Creek in Wet, only TN and TP were consistently lower in Pond1 (sediment basin) compared to stormwater inflow concentrations. Whilst there was considerable variation in nutrient concentrations between Ponds2–5, including increases in NH₄-N, by the time the stormwater reached Pond6 outlet, there had been an overall reduction compared to Pond1 outflow especially in NO_x-N (78%) and PO₄-P (67%).

In Dry there was removal in Pond1, with the exception of NH₄-N. NH₄-N (also reflected in TN) increased in Pond2 and although reductions did occur; outflow concentrations at P6 were greater than stormwater inflows, so the system exported NH₄-N. Concentrations of other nutrients were reduced, with removal of up to 90% NO_x-N and PO₄-P, compared to stormwater inflows. Biological uptake by plants, algae and phytoplankton would have played a major role, especially given detention times of 20 days plus during dry periods.

3.1.3. Phytoplankton

Chlorophyll-a. At Golden Pond, chlorophyll-a was only measured in Dry base-flow periods and values were low (3.5 \pm 0.6 $\mu\text{g L}^{-1}$ in the sediment basin, 5.5 \pm 3.2 $\mu\text{g L}^{-1}$ in Wetland

Table 8
Chlorophyll-a ($\mu\text{g L}^{-1}$) an indicator of phytoplankton biomass in the Bridgewater Creek System. Dry (base flow) and Wet (storm events). Mean \pm SD.

	Pond 1 (sediment basin)	Pond 2	Pond 3	Pond 4	Pond 5	Pond 6
Dry	64 \pm 80	54 \pm 60	15 \pm 20	12 \pm 12	10 \pm 5	12 \pm 10
Wet	12 \pm 15	22 \pm 30	25 \pm 30	22 \pm 28	28 \pm 22	33 \pm 20

1, and $3.2 \pm 0.8 \mu\text{g L}^{-1}$ in Wetland 2). All values are below WQO ($8 \mu\text{g/L}^{-1}$). This 'wetland' system dominated by aquatic vegetation and less open water did not produce algal blooms.

The Bridgewater Creek 'pond' system with larger open water areas was prone to algal blooms. Huge variations in chlorophyll-a occurred in both Dry and Wet (Table 8); mean values exceeded WQO's. Large algal blooms periodically occurred in Ponds 1 and 2 during Dry, but chlorophyll-a was reduced in Ponds 3 to 6. Numerous microcystaceans, in particular cladocerans, found in Ponds 2 to 6, would have been active predators on the phytoplankton.

In Wet (storm event) lower chlorophyll-a values in Pond1 probably reflect dilution, whereas the higher chlorophyll-a in Ponds 2 to 6 compared with Pond1, appear to be a flushing-out effect.

3.2. Macroinvertebrates and mosquitoes

Both systems had greater macroinvertebrate species richness compared to the upstream concrete channels (Table 9). The constructed stormwater wetlands at Golden Pond had even more taxa than the natural Kameruka Wetlands downstream. The Bridgewater Creek stormwater ponds recorded 86 species over a 4 year period.

Neither constructed stormwater treatment system supported successful mosquito breeding. Table 10 shows less than 5% of dips over a 12 month period conducted during Year 2 contained mosquito larvae and when present, they were in very low numbers. Larvae generally occurred amongst dead vegetation and most were only the very juvenile 1st and 2nd instars. No pupae were found indicating that the larvae did not complete their lifecycle. Microcystaceans, dragonfly nymphs, hemipterans (water bugs, water boatman and surface striders), water beetle larvae and adults, are known predators of mosquito larvae (Greenway et al., 2003).

4. Discussion

4.1. Water quality

The reduction of suspended solids in a wetland due to settling shows an exponential development over time – first order kinetic decay model ($k-C^*$ model) (Kadlec and Wallace, 2009). The amount of 'decay' depends upon the retention time within the wetland or pond, however this never reaches zero due to 'background' or irreducible concentrations (C^*). Similarly reductions in nitrogen and phosphorus can be explained by the $k-C^*$ model. Kadlec and Wallace (2009) have identified irreducible concentrations for a range of contaminants.

Kasper and Jenkins (2007) report on an extensive daily sampling program at Bridgewater Creek wetland. Following a storm event significant reductions in TSS occurred within 24 h however it took 72 h for TSS to decay to C^* . TSS in the outlet Pond6 was only lower than in the sediment basin Pond1 during storm events, confirming results presented in this paper. They attributed increases in TSS in dry weather to 3 factors: poor design of the inlet structure between Pond1 and Pond2 causing a zone of recirculation and turbulence; lack of macrophytes (aquatic vegetation) in Ponds2, 3, 4 reducing filtration/settlement and sediment stabilisation (Greenway, 2004); water bird activity causing resuspension in Ponds2, 3, 4. Greenway-this paper (Table 8) also found high chlorophyll-a in Pond2 which would contribute to TVS.

Lack of macrophytes can be related to both poor design and lack of maintenance (Greenway et al., 2007; Hunt et al., 2011). Water bird activity can be related to anthropogenic influence since local residents and visitors commonly feed the ducks and ibis from the board walk located between Ponds2, 3, 4. The viewing platform and benches overlooking Pond3 is a particularly popular spot for bird feeding as observed by Kasper and Jenkins (2007) and Greenway.

Table 9
Macroinvertebrate taxa in Golden Pond and Bridgewater Creek systems. *Indicates taxa that are predators of mosquito larvae.

Macroinvertebrate taxa	Golden pond			Bridgewater Creek	
	Concrete channel upstream	Stormwater wetlands 1&2	Natural wetland downstream	Concrete channel upstream	Stormwater ponds 1-6
Annelida, Nematoda, Flatworms ('worms')	3	6	6	3	9
Gastropoda (water snails)	5	8	8	4	5
Copepoda, Ostracoda, Cladocera*(microcystaceans)	3	5	4	1	8
Acarina*(water mites,spiders)	3	5	1		6
Epiroctophora* (dragonflies)	5	11	5	6	6
Zygoptera (damselflies)	1	5	5	1	5
Ephemeroptera (mayflies)	1	1	1	1	1
Hemiptera* (water bugs, water boatmen, striders)	1	6	2	2	15
Diptera (flies, mosquitoes)	2	3	2	6	10
Coleoptera*(water beetles)	1	4	4	3	19
Trichoptera (caddis flies)		1	1		2
TOTAL TAXA	25	55	39	27	86

Table 10
Relative abundance of mosquito larvae at Golden Pond (Wetlands 1 &2) and Bridgewater Creek (Ponds 1–6) in Year 2 following construction.

Location	Number of dips	% With larvae	% <10 larvae	% >1–40 larvae	% Pupae
Golden Pond	130	4.8	4.0	0.8	0
Bridgewater Creek	220	5.0	4.6	0.4	0

Table 11
Mean values for C* ('irreducible concentrations') for Dry weather at Outlet Pond1 (Sediment Basin) and Outlet Pond6 at Bridgewater Creek Stormwater Treatment System. (Source; Kasper and Jenkins, 2007). Compared to this study and Event Mean Concentrations: EMC (GCWWE, 2008) and Water Quality Objectives: WQO (BCC, 2000).

	TSS mg/L	TVS mg/L	TN mg/L	NH ₄	NO _x	TP mg/L	PO ₄
Pond 1 out	8.99*	10	6.61*	7	1.255*	1.28	na*
Pond 6 out	13.69*	16	6.00*	7	1.041*	1.04	na*
EMC: GCWWE 2008	19	na	1.21	na	0.20	0.58	na
WQO:BCC 2000	15	na	0.65	0.35	0.13	0.07	0.35

Signs have now been erected to advise the public "Do Not Feed The Birds".

Kasper and Jenkins (2007) determined C* for suspended solid and nutrients in the different ponds. Table 11 presents the Pond1 and Pond6 outlet values. Hathaway and Hunt (2010) summarised data for outflow 'irreducible concentrations' for storm water wetland basins produced from the International Stormwater BMP Database by Geosyntec Consultants and Wright water Engineers, Inc (GCWWE, 2008). The data for both Golden Pond and Bridgewater Creek are within these values for dry weather, confirming 'irreducible concentrations' ie any further reductions in TSS, TN and TP could not be expected. In storm events only TSS exceeded these 'irreducible concentrations'.

However the Water Quality Objectives (WQO) proposed by Brisbane City Council (BCC, 2000) for stormwater quality are very stringent and fail to recognise the concept of 'irreducible concentrations' for nutrients in wetland systems. Thus, it is not surprising that neither the Golden Pond nor the Bridgewater Creek stormwater wetland treatment systems met the desired WQO's for nutrients.

Hathaway and Hunt (2010) investigated the performance of 3 stormwater wetland cells (total 0.32 ha) in series in a small 12.5 ha watershed in North Carolina USA. For the storm events monitored final mean outlet 'irreducible' concentrations were 9 mg/L TSS; 0.07 mg/L NO_x; 0.72 mg/L TN; and 0.10 mg/L TP similar to Golden Pond and Bridgewater Creek with the exception of TSS. Hathaway and Hunt also investigated nitrogen speciation and noted an increase in Total Kjeldahl Nitrogen (ammonium and organic nitrogen ON) between Cell 2 and Cell3 which they attributed to organic N with an 'irreducible' concentration of 0.65 mg/L. At Golden Pond and Bridgewater Creek outlets ON was 0.58 and 0.61 mg/L respectively ie similar 'irreducible' concentration. However ON was very high 2.10 mg/L at Pond1 outlet and progressively reduced in concentration as the stormwater flowed through the wetland ponds.

Increases in NH₄-N within wetlands can be attributed to the ammonification of accumulated *in-situ* organic matter coupled with lower sediment interface oxygen, thereby limiting nitrification (Kadlec and Wallace, 2009). Bayley (2007) also reported export loads of NH₄-N between Pond1 and Pond6 at Bridgewater Creek wetland. During the winter dry months he noted that NH₄-N dominated the TN pool which he attributed to mineralisation of Organic N; lower redox potential (limiting nitrification) and a lower phytoplankton biomass (limiting biological uptake). Hathaway and Hunt (2010) recorded increased organic N and TKN in Wetland Cell 3 indicating *in-situ* release of NH₄-N.

Table 12
Social attributes and opportunities associated with the Golden Pond and Bridgewater Creek Constructed Stormwater Wetland Systems.

Attribute/opportunity	Golden pond system	Bridgewater Creek system: Bowie's Flat wetland
Recreation	Walking trails; linkages; bird watching	Walking trails; boardwalks; linkages; observation decks; bird watching
Education Research	Interpretive signage; website	Interpretive signage; brochures; websites; used by schools; universities
Landscape	Open water; aquatic plants; terrestrial vegetation; picnic table/shelter; car parking	Open water; aquatic plants; terrestrial vegetation; seats/benches; picnic tables/shelters; car parking water storage/irrigation
Community Engagement		Community catchment group; community working bees; educational workshops

Moore et al. (2011) compared ON and TN fractions and removal in 7 stormwater wetlands in North Carolina and found ON outlet concentrations ranged from 0.7 to 0.8 mg/L. They found no distinction between dry and wet weather samples. In both North Carolina and Brisbane wetlands ON outlet concentrations were significantly less than inflow stormwater runoff, thus despite irreducible concentrations, constructed stormwater wetlands are effective in reducing ON exports downstream. These 'irreducible' concentrations for organic N in stormwater wetland outlets, gives further credence to the rather stringent Water Quality Objective of 0.65 mg/L TN in the BCC (2000) WQO guidelines.

Outflow ON: TN ratios ranged from 0.33 to 0.98 with a median value of 0.75, again comparable to the 2 Brisbane wetlands. However the mean ON: TN ratio for stormwater runoff was 0.65 which differed from Golden Pond 0.26 and Bridgewater Creek 0.85 in storm events. In dry events only 2% of TN was ON (93% being NO_x) flowing into Golden Pond. The range of values reflects differences in catchment characteristics.

Moore et al. (2011) suggest that 'threshold effluent concentrations' should be considered as a more meaningful metric for nitrogen performance since % removal does not account for 'irreducible' background concentrations.

4.2. Macroinvertebrates

The creation of these stormwater wetlands from concrete drainage channels has provided habitats for a wide range of macroinvertebrates. Although not quantitatively measured, other wild life associated with both systems included frogs, reptiles and birds.

Herrmann (2012) monitored macroinvertebrate colonisation over 3 years in a 1 ha constructed stormwater wetland in Kalmar, Sweden. The highest number of taxa (55 species) was found in year1 with the greatest diversity of Coleoptera (22 spp) and Hemiptera (11 spp), but taxa decreased to 36 with Diptera (9 spp) having the greatest diversity. These changes in community composition were attributed to the early colonisation ability of beetles and corixids into the more open water body. At Bridgewater Creek, Coleoptera (19 spp) and Hemiptera (15 spp) as well as micro-crustaceans, were also the most diverse group due to the large expanse of open water in the ponds. However dense littoral vegetation also provided a habitat for many taxa, resulting in an increase in total taxa over the 4 years (Greenway and Polson, 2007). Thus the Bridgewater Creek pond system provided both open water and

vegetated habitats compared to the more densely vegetated wetland at Golden Pond.

Moore and Hunt (2012) quantitatively assessed macroinvertebrates in 20 stormwater wetland and 20 stormwater ponds in North Carolina, USA. They quantified biodiversity in terms of species richness (families) and found that wetlands and ponds supported similar diversity but found differences in community composition. Wetlands, and ponds with 'vegetated shelves' (littoral vegetation), supported a greater proportion of mosquito predators and a lower frequency of mosquito larvae. These findings support this study and the previous work of Greenway et al. (2003), Greenway and Polson (2007) which identified the importance of aquatic plant habitat for macroinvertebrate diversity. With habitat, rather than water quality, influencing total species richness.

4.3. Social values

The construction of stormwater treatment wetlands in these existing residential suburbs by retrofitting the concrete drains has enhanced social services and the landscape amenity by providing blue and green open space (Table 12). Both wetlands offer a range of recreational and educational attributes and opportunities, however the Bridgewater Creek/Bowie's Flat Wetland provides the greatest opportunities. Brisbane City Council, the Norman Creek Catchment Coordinating Committee, bird watching and bushwalking community groups all promote activities. Additionally the rainwater runoff 'stored' in the wetlands provides an opportunity for stormwater harvesting and is reused to irrigate the surrounding landscape.

Moore and Hunt (2012) qualitatively assessed 'cultural services' of stormwater wetlands and ponds using a rubric framework with scores for similar social attributes. Based on their scoring criteria, Bridgewater Creek would achieve the highest score 4/4 for recreation and education and Golden Pond 1/4. The wide range of social benefits at the Bridgewater Creek: Bowie's Flat Wetland can be attributed to community consultation in the planning stage involving Brisbane City Council, the Bridgewater Creek Community Reference Group and the Norman Creek Catchment Coordination Committee. In recognition, Brisbane City Council received a Healthy Waterways Government Award for the 'Bridgewater Creek Water Quality Improvement Project' (Water by Design-Bridgewater Creek Wetland).

5. Conclusions

These 2 retrofit stormwater wetlands constructed from concrete drainage channels improved water quality and created aquatic biodiversity in an urbanised setting. Both wetlands had a similar catchment size (180 ha) however water quality monitoring demonstrated that the larger 0.8 ha Bridgewater Creek system with a 12,500 m³ storage capacity was more effective in reducing suspended solids during storm events. This was also attributed to the design which incorporated an overflow bypass channel. At Golden Pond high flow-through velocities caused by the v-notch weir prevented settling and resuspended sediment. During dry weather duck activity in both wetlands caused resuspension of sediment. Both wetlands were effective in removing nutrients; however outflow concentrations did not meet Water Quality Objectives due to 'irreducible' background concentrations. This suggests that Brisbane City Council's stormwater quality objectives are too stringent for constructed wetland stormwater treatment devices.

Monitoring water quality in the natural Kameruka Wetlands demonstrated the importance of retaining and incorporating any existing wetlands as a component of a stormwater treatment chain.

Open water and dense littoral vegetation provided macroinvertebrate habitats. Macroinvertebrate assemblages in the constructed wetlands were higher than the concrete drainage channels; and even higher than in the Kameruka Wetlands.

Both constructed wetland systems provided green and blue infrastructure not only to improve stormwater quality but to create an environment for nature and people. Community and stakeholder engagement in the planning stage of the Bridgewater Creek wetland ensured the greatest opportunities; transforming a featureless open space dissected by a concrete drain into a wildlife haven and a major community attraction.

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