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**Published**

2008

**Journal Title**

Molecular Ecology

**DOI**

[10.1111/j.1365-294X.2007.03637.x](https://doi.org/10.1111/j.1365-294X.2007.03637.x)

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Molecular evidence for sequential colonization and taxon cycling in freshwater decapod shrimps on a Caribbean island

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Key words: amphidromy, nested clade analysis, population expansion, Puerto Rico

10 Running title: Taxon cycling in Caribbean freshwater shrimp

## Abstract

Taxon cycling, i.e. sequential phases of expansions and contractions in species' distributions associated with ecological or morphological shifts, are postulated to characterize dynamic biogeographic histories in various island faunas. The Caribbean freshwater shrimp assemblage is mostly widespread and sympatric throughout the region, although one species (*Atyidae: Atya lanipes*) is geographically restricted and ecologically and morphologically differentiated from other *Atya*. Using patterns of nucleotide variation at the COI mtDNA gene in five species of freshwater shrimp (*A. lanipes*, *A. scabra*, *A. innocuous*; *Xiphocarididae: Xiphocaris elongata*; *Palaemonidae: Macrobrachium faustinum*) from Puerto Rico, we expected to detect a signature of sequential colonization in these shrimp, consistent with the concept of taxon cycling, and expected that *A. lanipes* would be at a different taxon stage (i.e. an early stage species) to all other species. We also examined patterns of genetic population structure in each species expected with poor, intermediate and well-developed abilities for among river dispersal. Population expansions were detected in all species, although the relative timing of the expansions varied among them. Assuming that population expansions followed colonization of Puerto Rico by freshwater shrimp, results bear the hallmarks of sequential colonisation and taxon cycling in this fauna. *A. lanipes* had a star phylogeny, low mean pairwise nucleotide differences, and recent (Holocene) estimates for an *in situ* population expansion in Puerto Rico, and was inferred as an early stage species in the taxon cycle undergoing a secondary phase of expansion. All other species were inferred as late stage species undergoing regional population expansions, as their mean pairwise nucleotide differences were relatively high and phylogenetic patterns were more complex than *A.*

*lanipes*. High rates of gene flow without isolation-by-distance among rivers were detected in all species, although results should be treated cautiously as some populations are unlikely to be in mutation-drift equilibrium. Nested clade analysis produced inconsistent results among species that all have high rates of gene flow and expanding populations.

40

## Introduction

Oceanic islands have served as templates for studying biogeographic phenomena, such as area- and isolation- diversity relationships and strategies for colonization and biotic exchange (MacArthur & Wilson 1967). The concept of taxon cycling, and the similar  
45 concept of taxon pulses, is central in island biogeography (Darlington 1943; Wilson 1959, 1961; Erwin 1981; Liebherr & Hajek 1990), although it is not a universally accepted process (e.g. Pregill & Olsen 1981). The former proposes that taxa undergo sequential (although not unidirectional) phases of expansion and contraction, whether in distribution and/or ecological relationships (Ricklefs & Bermingham 2002), whereas the  
50 latter is a unidirectional model of distributional or ecological shift that facilitates speciation (Liebherr & Hajek 1990). Taxon cycles and taxon pulses are thus mechanisms of diversification over ecological and evolutionary timescales, respectively (Liebherr & Hajek 1990). Traditional approaches used to infer taxon cycles involved establishing relationships between distributional patterns and phenotypic or ecological differentiation  
55 (Wilson 1959, 1961). Thus, in the presence of some degree of phenotypic and/or ecological differentiation, an expanding species (i.e. a “late-stage” species in the taxon cycle) would have a widespread and continuous distribution, whilst a contracting species (i.e. an “early-stage” species) would have a restricted or fragmented distribution (Ricklefs & Bermingham 2002). Contracting species (i.e. early stage species) often undergo  
60 ecological shifts into marginal habitats effected by competition from colonizing late-stage species and may initiate secondary phases of expansion after adapting to new habitat types or experiencing ecological release from predators or disease (Wilson 1961; Erwin 1981; Ricklefs & Bermingham 2002).

65 As taxon cycles are historical processes, the phylogenetic information contained in  
mtDNA sequences makes molecular phylogeography an additional and highly  
informative approach to studying taxon cycles in island biota (Ricklefs & Bermingham  
1999; Emerson 2002). Comparative phylogeographic studies have often focused on  
among-island phylogeographic patterns to infer taxon cycles (e.g. Seutin *et al.* 1994;  
70 Lovette *et al.* 1998). However, the characteristic signatures that colonization and  
population expansion events leave in contemporary patterns of genetic variation  
(Templeton *et al.* 1995; Gübitz *et al.* 2005) mean that within-island phylogeographic  
patterns can be used to elucidate differences in the timing of population expansion among  
co-distributed taxa and develop putative scenarios for the chronology of colonization and  
75 temporal changes in assemblage structure (Brown & Pestano 1998; Vangergast *et al.*  
2004; Gübitz *et al.* 2005). For example, star-like phylogenetic patterns, low pairwise  
nucleotide differences in DNA sequence data, and molecular clock estimates for recent  
expansions would be indicative of a recent and perhaps localized expansion. This genetic  
pattern could reflect an early stage species that is undergoing a secondary expansion if it  
80 is ecologically divergent and has a restricted or fragmented distribution. In contrast,  
complex phylogenetic patterns, larger pairwise nucleotide differences in DNA sequence  
data, and molecular clock estimates for older population expansions likely reflect older  
and/or multiple expansions from multiple places, characteristic of late stage species.  
These phylogeographic expectations for expanding or recently expanding populations  
85 provide testable hypotheses that can be reconciled with distributional and ecological data  
to explore prospects for taxon cycling in island biota.

The freshwater fauna of the Caribbean archipelago is dominated by migratory species of caridean shrimp (e.g. genera *Atya*, *Xiphocaris*, *Macrobrachium*). These species are all  
90 amphidromous, meaning that their larvae are released into freshwater which then passively drift to estuarine or marine habitats and migrate upstream as post-larvae to headwater adult habitats (McDowall 2007). However, the extent to which amphidromy facilitates dispersal among rivers within an island (or at other spatial scales, e.g. among islands) and the length of time larvae remain in marine littoral areas is unknown  
95 (Holmquist *et al.* 1998). Despite this, the dependency of these shrimp on marine habitats as larvae has facilitated mostly widespread and sympatric distributions throughout the Caribbean archipelago, with the exception of *Atya lanipes* which is restricted to the Greater Antilles and Virgin Islands (Hobbs & Hart 1982; Fièvet 1998). Furthermore, different genera of these shrimps occupy semi-discrete ecological niches (i.e. *Atya* – filter  
100 feeders and scrapers; *Xiphocaris* – shredders; *Macrobrachium* – predatory and omnivorous; Crowl *et al.* 2001) and within the genus *Atya*, different species have distinct habitat preferences (e.g. *A. lanipes* – slow flowing (pool) habitats; *A. scabra* – fast flowing (riffle) habitats; Chace & Hobbs 1969). Interestingly, the filter feeding niche of *Atya* is most effective in riffle (fast flowing) habitats (Fryer 1977), suggesting that slow  
105 flowing pools inhabited by *A. lanipes* may be marginal habitat. Furthermore, *A. lanipes* is the most morphologically distinct *Atya* (Hobbs & Hart 1982) and it is not a recently derived taxon in the molecular phylogeny of the genus (Page *et al.* in press).

The varying patterns of distribution in these shrimp, their well-developed abilities for  
110 range expansion over evolutionary timescales and various levels of ecological  
differentiation lead us to hypothesize that taxon cycling characterises their biogeographic  
history. Using COI mtDNA variation in five species of amphidromous shrimp (Atyidae:  
*Atya lanipes*, *A. scabra*, *A. innocuous*; Xiphocarididae: *Xiphocaris elongata*;  
Palaemonidae: *Macrobrachium faustinum*) from Puerto Rico, we tested this hypothesis  
115 and expected to detect a distinct chronology in timing of population expansions in each  
species. In particular, we expected that *A. lanipes* would bear the molecular signatures of  
an early stage species, perhaps undergoing a secondary expansion, and that the other  
species would likely have molecular signatures expected for late stage species in the  
taxon cycle. Finally, as some Puerto Rican populations of caridean shrimp have declined  
120 or been locally extirpated in response to anthropogenic disturbance (e.g. dam  
construction, water abstraction and harvest-related poisoning) (Benstead *et al.* 1999;  
Greathouse *et al.* 2005), and they are known to have important functions in Caribbean  
island stream ecology (Pringle *et al.* 1993, 1999; Crowl *et al.* 2001), knowledge of  
within-island (among-river) dispersal in these shrimp may be important for river  
125 management (Holmquist *et al.* 1998). We predicted that if these shrimps could disperse  
effectively through marine habitats (at the island scale), we would detect no genetic sub-  
structure among populations from different rivers or marine regions in Puerto Rico and  
reveal no signatures of isolation-by-distance (IBD) in the genetic data. In contrast, if  
these shrimp had poor or intermediate abilities to disperse via marine habitats, we would  
130 find significant genetic sub-division among rivers or among marine regions, respectively,  
and detect significant patterns of IBD in the genetic data.



## Methods

### Sampling design and laboratory techniques

135 A hierarchical design was implemented to test the above predictions: three regions (i.e. Atlantic Ocean, Mona Passage, Caribbean Sea), each with three rivers (Fig. 1). Sample sizes are shown in Table 1. Genomic DNA was extracted from each individual using a standard phenol-chloroform procedure, and a fragment of the cytochrome *c* oxidase subunit 1 (COI) mtDNA gene was amplified via PCR using specifically developed

140 primers: CR-COI-F: CWACMAAYCATAAGAYATTGG; CR-COI-R: GCRGANGTRAARTARGCTCG. PCR reactions contained approximately 40 ng of template DNA, 0.4  $\mu$ M of each primer, 0.2 mM dNTP (Astral Scientific), 2 mM MgCl<sub>2</sub>, 1.25  $\mu$ L of 10x polymerase reaction buffer and 0.25 unit of *Taq* polymerase (Fisher Biotech), adjusted to a final volume of 12.5  $\mu$ L with ddH<sub>2</sub>O. The thermal-cycling profile

145 followed: 5 minutes at 94°C; 35 cycles of 30 seconds at 94°C , 30 seconds at 55°C and 30 seconds at 72°C; an additional extension phase of 5 minutes at 72°C; and a final hold stage at 4°C. PCR product was purified with the exonuclease I-shrimp alkaline phosphatase method, using 2.5  $\mu$ L PCR product, 2.0  $\mu$ L shrimp alkaline phosphatase (Promega) and 0.5  $\mu$ L exonuclease I (Fermenta), and a two-step thermal-cycling profile:

150 35 minutes at 37°C, 20 minutes at 80°C. Sequencing reactions contained 0.5  $\mu$ l purified product, 0.32  $\mu$ l forward primer, 2  $\mu$ l BigDye v1.1 (Applied Biosystems) and 2  $\mu$ l 5x sequencing buffer (Applied Biosystems), and the following thermal cycling conditions were used: 1 minute at 96°C; 30 cycles of 10 seconds at 96°C, 5 seconds at 50°C, 4 minutes at 60°C; and a hold period of 4°C. Sequencing was conducted on a 3130xl

155 Capillary Electrophoresis Genetic Analyzer (Applied Biosystems) and sequences were aligned and edited using SEQUENCHER version 4.1.2 (Gene Codes). An exemplar of each haplotype was sequenced in the reverse direction to verify bases at polymorphic sites.

#### 160 Data analysis

Haplotype ( $h$ ), nucleotide ( $\pi$ ) diversity and mean pairwise nucleotide differences ( $k$ ) were calculated for each species in ARLEQUIN (Schnieder *et al.* 2000) to obtain measures of molecular diversity in each species. The parameters  $D$  (Tajima 1989),  $F_s$  (Fu 1997), and  $R_2$  (Ramos-Onsins & Rozas 2002) were calculated in DnaSP (Rozas *et al.* 2003) to  
165 examine signatures of population expansions in the mtDNA data. Their significance was assessed using 10,000 coalescent simulations, given the observed number of segregating sites. Estimates for  $\tau$ , including lower- and upper- bound estimates, were also calculated for each species in ARLEQUIN and used to determine the number of generations since the last population expansion using the formula of Rogers & Harpending (1992) and  
170 assuming a sequence evolution rate of 1.4% per million years (Knowlton & Weight, 1998). Fragment lengths for each species are presented in Table 1 and a generation time of two years was assumed for each species as their growth rates are slow in comparison with caridean shrimp in other parts of the world (*c.f.* Yam & Dudgeon 2005) and protandry is reported in some species of *Atya* (Carpenter 1978). AMOVA (Excoffier *et al.* 1992) was then implemented using ARLEQUIN to examine spatial mtDNA variation  
175 in each species according to the hierarchal design, and the relationship between mtDNA divergence among populations and geographic distance was tested using Mantel tests

(Mantel 1967) in PRIMER version 5.2.8 (Clarke & Gorley 2001). Finally, haplotype networks were constructed using TCS version 1.18 (Clement *et al.* 2000) and cladistic analysis of the nested haplotypes (NCA, Templeton *et al.* 1995) was implemented using  
180 GeoDis version 2.4 (Posada *et al.* 2000). The November 2005 inference key was used in NCA (available at: Darwin.uvigo.ed/download/geodisKey\_11Nov05.pdf).

## Results

185 Genetic diversity and mean pairwise nucleotide differences were much lower in *A. lanipes* than all other species (Table 1). Tajima's *D* and Fu's *F<sub>s</sub>* were negative in all species (Table 2), with *F<sub>s</sub>* being significant for all taxa and *D* being significant for all except *A. innocuous*. *R<sub>2</sub>* indicated significant signatures of population expansions in all amphidromous shrimp except *A. innocuous*. The number of generations that has elapsed  
190 since the last population expansion differs substantially among some of the taxa, although the variance around these estimates is large (Fig. 2). The relative timing of population expansions for various species-pair combinations were non-overlapping (Fig. 2), although were indistinguishable among some of the species (i.e. no differences among *A. lanipes*, *A. scabra* and *M. faustinum*). Most noteworthy are the non-overlapping estimates for *A.*  
195 *lanipes* and *A. innocuous*, as they are congeners - *A. innocuus* had an upper bound limit for a population expansion that coincided with the late Pleistocene (approximately 50,000 years BP), whilst *A. lanipes* had an upper bound limit for a late Holocene population expansion (approximately 10,000 years BP; Fig. 2). AMOVA revealed that genetic variation was not significantly partitioned either among regions or among rivers within  
200 regions for the amphidromous shrimp ( $\Phi_{CT}$ ,  $\Phi_{SC}$  and  $\Phi_{ST}$  were non-significant for all

taxa), with the exception of *M. faustinum*, which had significant genetic differentiation among rivers within regions ( $\Phi_{SC} = 0.040$ ,  $P = 0.027$ ). Pairwise analyses of  $\Phi_{ST}$  among rivers for the amphidromous shrimp found no significant differences after the Bonferroni correction for multiple tests was applied (data not shown). There were no significant  
205 patterns of isolation-by-distance in any of the amphidromous shrimp for shortest coastline distance among river mouths (*A. lanipes*,  $\rho = 0.298$ ,  $P = 0.096$ ; *A. scabra*,  $\rho = 0.072$ ,  $P = 0.335$ ; *A. innocuous*,  $\rho = 0.500$ ,  $P = 0.498$ ; *X. elongata*,  $\rho = -0.943$ ,  $P = 1.000$ ; *M. faustinum*,  $\rho = 0.073$ ,  $P = 0.409$ ) nor straight line geographic distance among river  
210 mouths (*A. lanipes*,  $\rho = -0.049$ ,  $P = 0.597$ ; *A. scabra*,  $\rho = 0.076$ ,  $P = 0.291$ ; *A. innocuous*,  $\rho = -0.500$ ,  $P = 0.834$ ; *X. elongata*,  $\rho = 0.464$ ,  $P = 0.206$ ; *M. faustinum*,  $\rho = 0.497$ ,  $P = 0.600$ ). The genealogical patterns revealed in Puerto Rican freshwater shrimp varied considerably among species, including among congeners of *Atya* (Fig. 3). NCA revealed contiguous range expansion in some taxa and restricted gene flow with isolation-by-distance in other species (Table 3).

215

## Discussion

### *Sequential colonization and taxon cycling*

We expected to detect a distinct chronology in the timing of population expansions in amphidromous shrimp in Puerto Rico. Our results indicate that population expansions in  
220 *A. innocuous* and *X. elongata* preceded those in *M. faustinum* and *A. lanipes*, suggesting that the freshwater shrimp assemblage structure has changed over the recent past, assuming that expansions were associated with a founder event. Differential colonization of islands is reported to facilitate the development of a temporally dynamic assemblage

structure in birds, rodents and lizards (Roughgarden & Pacala 1989; McFarlane &  
225 Lundberg 2002; Ricklefs & Bermingham 1999) and our data supports ideas that time is  
an important predictor of species richness on islands (Parent & Crespi 2006). The  
chronology of colonization in Puerto Rican amphidromous shrimp, coupled with  
demographically growing populations, also supports the concept of taxon cycling and that  
island biotic assemblages and population sizes are changing in response to ecological  
230 and/or evolutionary drivers.

In attempting to characterize the taxon cycle for Puerto Rican shrimp, we predicted that  
*A. lanipes* would be an early stage species on account of its restricted distribution and  
morphological and ecological distinctiveness (Fryer 1977; Hobbs and Hart 1982). This  
235 prediction was supported by the genetic data, as the haplotype network reflected a star  
phylogeny, mean pairwise nucleotide diversity was low, molecular clock estimates for the  
expansion were recent, and NCA revealed contiguous, and most likely *in situ*, range  
expansion. On their own, these genetic patterns could also have suggested that *A. lanipes*  
is a late stage taxon at the very start of an expansion phase. However, the known  
240 distinctiveness of *A. lanipes* in terms of ecology, morphology and phylogenetic  
relationships with other *Atya* suggest that it is not a recently derived taxon; thus, a  
secondary expansion phase is a more likely interpretation. Furthermore, all Caribbean  
amphidromous shrimp are described as “littoral” species on the basis of their well  
developed abilities for dispersal and range expansion (Fièvet 1998), suggesting that  
245 distributional differences between *A. lanipes* and other shrimp are due to taxon cycling  
stages rather than differences in intrinsic dispersal abilities. Multiple colonisation events

among islands are shown to produce stronger founder effects (i.e. genetic impoverishment) than a single colonization from the original (mainland) population (Clegg *et al.* 2002). Very low levels of genetic diversity in *A. lanipes* suggest that its recent colonization of Puerto Rico may be due to sequential colonization events among-islands, and that colonization of Puerto Rico by *A. lanipes* was effected by dispersal from an already genetically impoverished population undergoing a secondary expansion. This idea would need formal examination by a broader phylogeographic study throughout the Greater Antilles and Virgin Islands.

255

Our prediction that *A. lanipes* is an early stage species in the taxon cycle is further supported by genetic data for the other shrimp species, which suggest that they are late stage species (i.e. more complex haplotype networks with higher mean pairwise nucleotide diversity). However, as there was a chronology of colonization events, even among these late-stage species (*c.f.* *A. innocuous* & *X. elongata* versus *M. faustinum*), results also indicate a continuum between early-stage and late-stage species in taxon cycles. The complex haplotype networks of the widely distributed shrimp (i.e. *A. scabra*, *A. innocuous*, *X. elongata* and *M. faustinum*) each had several divergent clades, with each clade containing several to many closely related haplotypes, suggesting the possibility for multiple expansions, most of which likely occurred outside of Puerto Rico. These species are thus likely to be experiencing regional population growth. NCA indicated continuous range expansion in *M. faustinum*, which is consistent with an expansion phase in the taxon cycle, whilst NCA indicated restricted gene flow with IBD for *A. scabra* and *X. elongata*. This result is not expected for expanding populations, and contrasts with results

270 of the Mantel tests and population demographic analyses, as IBD usually reflects limited  
dispersal in a population at gene flow-drift equilibrium (Slatkin 1985). Other studies have  
shown that NCA can infer restricted gene flow with IBD under non-equilibrium  
conditions (Alexandrino *et al.* 2002; Kotlík & Berrebi 2007), and our study indicates  
inconsistent results for NCA among species that all have high rates of gene flow and  
275 expanding populations in drift-mutation disequilibrium.

Another interesting result is the distinct period of habitation of Puerto Rico by the  
congeners *A. lanipes* and *A. innocuous*, indicating a period of over 15,000 years of  
allopatry on the island. Most species are thought to arise in allopatry (Templeton 1980),  
280 and historical allopatry is reported to facilitate diversification in other caridean shrimp,  
with subsequent dispersal accounting for present-day sympatric distributions (Cook *et al.*  
2006). The concept of taxon pulses is similar to taxon cycles, although rather than  
reflecting distributional or ecological shifts over ecological time, taxon pulses result in  
speciation over evolutionary timescales (Liebherr & Hajek 1990). The molecular  
285 evidence for recent (i.e. late Pleistocene - Holocene) taxon cycles in Caribbean  
amphidromous shrimp may provide clues about longer-term processes of distributional  
change, periods of allopatry and diversification via taxon pulses in these shrimp further  
back in time (*c.f.* Miocene-Pliocene origin of most Caribbean atyid species, Page *et al.* in  
press). Various marine biogeographic breaks are reported throughout the Caribbean,  
290 including Mona Passage (Taylor & Hellberg 2006), which may also represent barriers to  
dispersal and range expansion in other species that utilize marine environments, such as  
amphidromous shrimp. Marine biogeographic breaks could have facilitated differential

historical distributions in shrimp, thereby catalyzing taxon pulses and allopatric  
speciation. Interestingly, taxon pulses are often accompanied by character displacement,  
295 with body size being a commonly displaced character among species of *Anolis* lizards  
(Losos 1992). The sister group to Caribbean atyid shrimps are large-bodied Indo-Pacific  
species, suggesting that speciation via taxon pulses in Caribbean atyids was also  
accompanied by size displacement, as evidenced by the radiation of small-bodied atyids  
in the region (Page *et al.* in press).

300

Identifying the drivers of assemblage structure change and taxon cycles on oceanic  
islands is a challenge and remains speculative (Ricklefs & Bermingham 2002). Using  
paleontological data for Caribbean vertebrates, Pregill & Olsen (1981) invoked  
Pleistocene aridity throughout the Caribbean, and associated shifts from xeric to mesic  
305 habitats and changing patterns of terrestrial island connectivity, to explain changes in  
assemblage structure and distribution in vertebrate fauna. Indeed they concluded cyclic  
changes in the distribution of Caribbean vertebrates did not characterize their  
biogeographic historic and went as far as to reject the taxon cycle as a nonexistent  
phenomenon. However, more recently, competition and host disease resistance were  
310 invoked as putative drivers of taxon cycles in Caribbean birds (Ricklefs & Bermingham  
1999, 2002), taxon cycles were shown for Antillean rodents (MacLean & Lundberg  
2002), and competition across habitat gradients was shown to produce temporally  
dynamic changes in distribution in Caribbean geckos (MacLean & Holt 1979); thus,  
taxon cycling can certainly not be rejected for Caribbean vertebrates. Although  
315 Pleistocene aridity in the Caribbean region may have facilitated evolution of adaptations



to pool habitats in *A. lanipes*, as aridity would have reduced the number of fast-flowing habitats in Caribbean island rivers, competition is also likely to have been a contributor to habitat specialization, as *A. scabra* has adaptations suited to riffles where filter feeding is more effective (Fryer 1977). Furthermore, Caribbean shrimp occupy semi-discrete  
320 feeding guilds (Crowl *et al.* 2001), suggesting ecological segregation over trophic niches. Thus, although low flow conditions greatly reduce shrimp abundance (Covich *et al.* 2003), and this would have been exacerbated during historical periods of aridity, competition across ecological gradients would also have had a role in changing distribution and taxon cycling in these shrimp. If changing climatic conditions was the  
325 sole factor facilitating population expansions in our study species, we would have expected to detect a strong degree of congruence in the relative timing of their population expansions associated with a pronounced climatic event. This may have also indicated population expansions in pre-established populations and led to the rejection of our prediction for sequential colonisation. However, population expansions were often non-  
330 contemporaneous and the absence of large historical climatic change in the Caribbean region throughout the Holocene (Metcalf *et al.* 2000) make it unlikely that these shrimp expanded from small pre-existing populations in response to a profound historical event. Taxon cycling in Caribbean birds is also reported to occur independently of historical climatic changes and occurs independently among species (Ricklefs & Bermingham  
335 2002). Disturbances (e.g. hurricanes and volcanoes) play important roles in determining Caribbean shrimp abundance and dynamic processes of species accumulation and turnover (Covich *et al.* 1991; Covich 2006). Thus, it is very likely that taxon cycling has

been a central process in the biogeographic history of Caribbean shrimp, driven by both ecological and environmental factors.

340

Although our data support taxon cycling in amphidromous Caribbean shrimp, factors other than stage in the taxon cycle may also explain their patterns of nucleotide variation, including selective sweeps, genetic hitchhiking and background selection. For example, *A. lanipes* may be in the early stages of recovering from a selective sweep, whereby a gene tightly linked with the COI mtDNA gene underwent intense selection, resulting in the population losing variation at the selected and linked genes. Similarly, the shrimp inferred as late stage species could have undergone selective sweeps further in the past, and their present-day high levels of genetic diversity could suggest that their populations have been recovering for longer. However, as noted earlier, there is strong ecological and biogeographic evidence for taxon cycling in Caribbean amphidromous shrimp, which explains patterns of nucleotide variation better than selection, genetic hitchhiking or background selection.

350

#### *Contemporary population genetic structure*

355

The amphidromous life history of Puerto Rican freshwater shrimp appears to facilitate effective genetic exchange among rivers, as shown on the basis of allozyme data for *A. lanipes*, suggesting that larval dispersal via the marine environment will likely lead to recolonisation of defaunated habitats if barriers to upstream migration are removed. This also suggests that the ecological functions of the shrimp assemblage (Pringle *et al.* 1993, 1999; Crowl *et al.* 2001) will be reinstated following changed river management

360

practices. Interestingly, Mona Passage represents a marine biogeographic break for various coral reef fishes between Puerto Rico and other islands of the Greater Antilles (Taylor & Hellberg 2006), although this water body appears not to restrict marine littoral dispersal of shrimp larvae among Puerto Rican rivers. In contrast, populations of volcano shrimp (*Halocaridina rubra*) on the island of Hawaii are reported to be genetically  
365 isolated among anchialine habitats (Santos 2006). This suggests that this species does not disperse via the ocean despite inhabiting mixohaline environments. Evidence for population expansions in all Puerto Rican shrimp species may mean that they are not in drift-mutation equilibrium and that the magnitude of gene flow could be overestimated,  
370 although larval dependency on marine littoral habitats suggests that very high gene flow is a biologically feasible interpretation. The significant  $\Phi_{SC}$  (i.e. “among rivers within regions”) for *M. faustinum* suggests that any one river (or pool in a river) may not reflect the overall genetic composition of the total population in this species, which is a pattern found in widespread fishes (Allendorf & Phelps 1981; Waples 1998) and widely-  
375 dispersing stream insects (Bunn & Hughes 1997). Degradation of riverine spawning habitat used by anadromous species (i.e. marine species that migrate to freshwater for reproduction, such as many salmonids) is reported to have serious consequences for the viability of populations as site fidelity greatly reduces opportunities for recolonisation among rivers (Policansky & Magnuson 1998; Jonsson *et al.* 1999; Prowles *et al.* 2000).  
380 In contrast, degradation of riverine habitat may have less-serious consequences for amphidromous species, as they appear to have effective capabilities for among-river recolonisation (McDowall 2003; Covich 2006; this study). However, multiple larval stages of Caribbean shrimp have been reported to stay in estuaries rather than drifting to

coastal marine waters (Benstead *et al.* 2000). This suggests that gene flow among rivers  
385 may be effected by a subset of larval recruits, perhaps by recruits from reproducing adults  
in reaches closer to estuaries, or occurs sporadically in association with hurricanes  
(Covich 2006). High rates of genetic exchange in several migratory riverine species have  
been shown to over-estimate abilities for among-river dispersal (Campana & Thorrold  
2001). Our results that show high genetic connectivity among rivers for Puerto Rican  
390 amphidromous shrimp should thus be treated with caution and river management should  
foster strategies that maximize within- and among- river recruitment.

## Acknowledgements

Research was funded by a Cooperative Agreement between the USDA Forest Service's  
395 International Institute of Tropical Forestry (IITF) and the University of Georgia. It was  
also supported by the National Science Foundation Luquillo LTER (Project DEB-  
0218039 and DEB-0620910) to IITF the University of Puerto Rico and the IITF USDA  
Forest Service, as part of the Long-Term Ecological Research Program in the Luquillo  
Experimental Forest. We thank the USDA Forest Service for permits to sample and the  
400 USD-FA and the Australian Rivers Institute, Griffith University, for financially  
supporting this research. Pablo Hernandez-Garcia and Katherine Smith helped during the  
sampling expedition, and Tim Page and three anonymous reviewers provided helpful  
comments on an earlier version of the manuscript. Daniel Schmidt helped design the  
primers we used.

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## References

- Alexandrino J, Arntzen JW, Ferrand N (2002) Nested clade analysis and the genetic evidence for population expansions in the phylogeography of the golden-stripped salamander, *Chioglossa lusitanica*. *Heredity*, **88**, 66-74.
- 410 Allendorf FW, Phelps SR (1981) Use of allelic frequencies to describe population structure. *Canadian Journal of Fisheries and Aquatic Sciences*, **38**, 1507-1514.
- Benstead JP, March JG, Pringle CM, Scatena FN (1999) Effects of a low-head dam and water abstraction on migratory tropical stream biota. *Ecological Applications*, **9**, 656-668.
- 415 Benstead JP, March JG, Pringle CM (2000) Estuarine larval development and upstream post-larval migration of freshwater shrimps in two tropical rivers of Puerto Rico. *Biotropica*, **32**, 545-548.
- Brown RP, Pestano J (1998) Phylogeography of skinks (Chalcides) in the Canary Islands inferred from mitochondrial DNA sequences. *Molecular Ecology*, **7**, 1183-1191.
- 420 Bunn SE, Hughes JM (1997) Dispersal and recruitment in streams: evidence from genetic studies. *Journal of the North American Benthological Society*, **16**, 338-346.
- Campana SE, Thorrold SR (2001) Otoliths, increments, and elements: keys to a comprehensive understanding of fish populations? *Canadian Journal of Fisheries and Aquatic Sciences*, **58**, 30-38.
- 425 Carpenter A (1978) Protandry in the freshwater shrimp, *Paratya curvirostris* (Heller, 1862) (Decapoda: Atyidae), with a review of the phenomenon and its significance in the Decapoda. *Journal of the Royal Society of New Zealand*, **3**, 343-358.

- Chace FA, Hobbs HH (1969) The freshwater and terrestrial decapod crustaceans of the  
430 West Indies with special reference to Dominica. *U.S. National Museum  
Bulletin*, **292**, 1-258.
- Clarke KR, Gorley RN (2001) *PRIMER v 5.2.8: User Manual/Tutorial*. Plymouth,  
PRIMER-E.
- Clegg SM, Degnen SM, Kikkawa J, Moritz C, Estoup A, Owens IPF (2002) Genetic  
435 consequences of sequential founder events by an island-colonizing bird.  
*Proceedings of the National Academy of Sciences*, **99**, 81-27-8131.
- Clement M, Posada D, Crandall K (2000) TCS: a computer program to estimate gene  
genealogies. *Molecular Ecology*, **9**, 1657-1660.
- Cook BD, Baker AM, Page TJ, Grant SC, Hurwood DA, Hughes JM (2006)  
440 Biogeographic history of an Australian freshwater shrimp, *Paratya australiensis*  
(Atyidae): the role life transition in phylogeographic diversification. *Molecular  
Ecology*, **15**, 1083-1093.
- Covich AP, Crowl TA, Johnson SL, Varza D, Certain M (1999) Post-hurricane Hugo  
increases in atyid shrimp abundances in a Puerto Rican montane stream.  
445 *Biotropica*, **23**, 448-454.
- Covich AP, Crowl TA, Scatena FN (2003) Effects of extreme low flows on freshwater  
shrimps in a perennial stream. *Freshwater Biology*, **48**, 1199-1206.
- Covich AP (2006) Dispersal-limited biodiversity of tropical insular streams. *Polish  
Journal of Ecology*, **54**, 523-547.
- 450 Crowl TA, McDowell WH, Covich AP, Johnson SL (2001) Freshwater shrimp effects on  
detrital processing and nutrients in a tropical headwater stream. *Ecology*, **82**, 775-

783.

Darlington PJ (1943) Carabidae of mountains and islands: data on the evolution of

isolated faunas, and on atrophy of wings. *Ecological Monographs*, **13**, 38-61.

455 Emerson BC (2002) Evolution on oceanic islands: molecular phylogenetic approaches to understanding pattern and process. *Molecular Ecology*, **11**, 951-966.

Erwin TC (1981) Taxon pulses, vicariance, and dispersal: an evolutionary synthesis illustrated by carabid beetles. In: *Vicariance biogeography: a critique* (eds. Nelson G, Rosen DE), pp. 159-196. Columbia University Press, New York

460 Excoffier L, Smouse PE, Quattro JM (1992) Analysis of molecular variance inferred from metric distances among DNA haplotypes: application to human mitochondrial DNA restriction data. *Genetics*, **131**, 479-491.

Fièvet E (1998) Distribution et capacités d'expansion des crevettes d'eau douce de la région caraïbe exemple des genres *Macrobrachium* et *Atya* (Crustacea: Caridea).

465 *Biogeographica*, **74**, 1-22.

Fryer G (1977) Studies on the functional morphology and ecology of the atyid prawns of Dominica. *Philosophical Transactions of the Royal Society B – Biological Sciences*, **277**, 57-129.

Fu YX (1997) Statistical tests of neutrality of mutations against population growth, 470 hitchhiking and background selection. *Genetics*, **147**, 915-925.

Greathouse EA, March JG, Pringle CM (2005) Recovery of a tropical stream after a harvest-related chlorine poisoning event. *Freshwater Biology*, **50**, 603-615.

Gübitz T, Thorpe RS, Malhotra A (2005) The dynamics of genetic and morphological



- variation on volcanic islands. *Proceedings of the Royal Society of London B*, **272**,  
475 751-757.
- Hobbs HH, Hart CW (1982) The shrimp genus *Atya* (Decapoda: Atyidae). *Smithsonian  
Contribution to Zoology*, **264**, 1-143.
- Holmquist JG, Schmidt-Gengenbach JM, Yoshioka BB (1998) High dams and marine-  
freshwater linkages: effects on native and introduced fauna of the Caribbean.  
480 *Conservation Biology*, **12**, 621-630.
- Jonsson B, Waples RS, Friedland KD (1999) Extinction considerations for diadromous  
fishes. *Journal of Marine Science*, **56**, 405-409.
- Knowlton N, Weigt LA (1998) New dates and new rates for divergence across the  
Isthmus of Panamá. *Proceedings of the Royal Society of London B*, **265**, 2257-  
485 2263.
- Kotlík P, Berrebi P (2007) Nested clade phylogeographical analysis of barbell (*Barbus  
barbus*) mitochondrial DNA variation. In: *Phylogeography of Southern European  
Refugia: evolutionary perspectives on the origins and conservation of European  
biodiversity* (eds, Weiss S & Ferrand N), pp. 315-325. Springer, The Netherlands.
- 490 Liebherr JK, Hajek AE (1990) A cladistic test of the taxon cycle and taxon pulse  
hypotheses. *Cladistics*, **6**, 39-59.
- Losos JB (1992) A critical comparison of the taxon cycle and character displacement  
models for size variation in *Anolis* lizards in the Lesser Antillers. *Copeia*, **1992**,  
279-288.
- 495 Lovette IJ, Bermingham E, Seutin G, Ricklefs EF (1998) Evolutionary differentiation in  
three endemic West Indian warblers. *Auk*, **115**, 890-903.

- MacArthur RH, Wilson EO (1967) *The Theory of Island Biogeography*. Princeton University Press, New Jersey.
- MacFarlane DA, Lundberg J (2002) A middle Pleistocene age and biogeography for the  
500 Extinct rodent *Megalomys curazensis* from Curacao, Netherlands Antilles. *Caribbean Journal of Science*, **38**, 278-281.
- MacLean WP, Holt RD (1979) Distributional patterns in St. Croix *Sphaerodactylus* lizards: the taxon cycle in action. *Biotropica*, **11**, 189-195.
- McDowall RM (2003) Hawaiian biogeography and the islands' freshwater fish fauna.  
505 *Journal of Biogeography*, **30**, 703-710.
- McDowall RM (2007) On amphidromy, a distinct form of diadromy in aquatic organisms. *Fish and Fisheries*, **8**, 1-13.
- Metcalf SE, O'Hara SC, Caballero M, Davies SJ (2000) Records of Late Pleistocene-Holocene climatic change in Mexico – a review. *Quaternary Science Reviews*, **19**,  
510 699-721.
- Page TJ, Cook BD, von Rintelen T, von Rintelen K, Hughes JM (in press) Evolutionary relationships of atyid shrimps imply both ancient Caribbean radiations and common marine dispersal. *Journal of the North American Benthological Society*.
- Parent CE, Crespi BJ (2006) Sequential colonization and diversification of Galapagos  
515 endemic land snail Genus *Bulimulus* (Gastropoda: Stylommatophora). *Evolution*, **60**, 2311-2328.
- Policansky D, Magnuson JJ (1998) Genetics, metapopulations, and ecosystem management of fisheries. *Ecological Applications*, **8**, S119-S123.
- Posada D, Crandall KA, Templeton AR (2000) GeoDis: a program for the cladistic

- 520 nested analysis of the geographical distribution of genetic haplotypes. *Molecular Ecology*, 9, 487-488.
- Pregill GK, Olsen SL (1981) Zoogeography of West Indian vertebrates in relation to Pleistocene climatic cycles. *Annual Review of Ecology and Systematics*, **12**, 75-98.
- 525 Pringle CM, Blake GA, Covich AP, Buzby KM, Finley A (1993) Effects of omnivorous shrimp in a montane tropical stream: sediment removal, disturbance of sessile invertebrates and enhancement of understory algal biomass. *Oecologia*, **3**, 1-11.
- Pringle CM, Hemphill N, McDowell WH, Bednarek A, March JG (1999) Linking species and ecosystems: different biotic assemblages cause interstream differences in  
530 organic matter. *Ecology*, **80**, 1860-1872.
- Prowles H, Bradford MJ, Bradford RG, Doubleday WG, Innes S, Levings CD (2000) Assessing and protecting endangered marine species. *Journal of Marine Science*, **57**, 669-676.
- Ramos-Onsins SE, Rozas J (2002) Statistical properties of new neutrality tests against  
535 population growth. *Molecular Biology and Evolution*, **19**, 2092-2100.
- Ricklefs RE, Bermingham E (1999) Taxon cycles in the Lesser Antillean avifauna. *Ostrich*, **70**, 49-59.
- Ricklefs RE, Bermingham E (2002) The concept of the taxon cycle in biogeography. *Global Ecology and Biogeography*, **11**, 353-361.
- 540 Rogers AR, Harpending H (1992) Population growth makes waves in the distribution of pairwise genetic differences. *Molecular Biology and Evolution*, 9, 552-569.
- Roughgarden J, Pacala S (1989) Taxon cycles among *Anolis* lizard populations: review of

- the evidence. In: *Speciation and its Consequences* (eds, Otte D, Endler JA), pp. 403-432. Sinauer Associates, Massachussets.
- 545 Rozas J, Sánchez-DelBarrio JC, Messeguer X, Rozas R (2003) DnaSP: DNA polymorphism analyses by the coalescent. *Molecular Biology and Evolution*, **9**, 552-569.
- Santos S (2006) Patterns of genetic connectivity among anchialine habitats: a case study of the endemic Hawaiian shrimp *Halocaridina rubra* on the island of Hawaii. *Molecular Ecology*, **15**, 2699-2718.
- 550 Schneider S, Kuffer J, Rossli D, Excoffier L (2000) *ARLEQUIN version 2.0: a software for population genetic data analysis*. Genetics and Biometry Laboratory, University of Geneva, Geneva.
- Seutin G, Klein NK, Ricklefs RE, Bermingham E (1994) Historical biogeography of the bananaquit (*Coereba flaveola*) in the Caribbean region: a mitochondrial DNA assessment. *Evolution*, **48**, 1041-1061.
- Slatkin M (1985) Gene flow in natural populations. *Annual Review of Ecology and Systematics*, **16**, 393-430.
- Tajima F (1989) Statistical method for testing the neutral mutation hypothesis by DNA polymorphism. *Genetics*, **123**, 585-595.
- 560 Taylor MS, Hellberg ME (2006) Comparative phylogeography in a genus of coral reef fishes: biogeographic and genetic concordance in the Caribbean. *Molecular Ecology*, **15**, 695-707.
- Templeton AR (1980) The theory of speciation via the founding principle. *Genetics*, **94**, 565 1011-1038.

- Templeton AR, Routman E, Phillips CA (1995) Separating population structure from population history: a cladistic analysis of the geographical distribution of mitochondrial DNA haplotypes in the tiger salamander, *Ambystoma tigrinum*. *Genetics*, **140**, 767-782.
- 570 Vandergast AG, Gillespie RG, Roderick GK (2004) Influence of volcanic activity on the population genetic structure of Hawaiian spiders: fragmentation, rapid population growth and the potential for accelerated evolution. *Molecular Ecology* **13**, 1729-1743.
- Waples RS (1998) Separating the wheat from the chaff: patterns of genetic differentiation  
575 in high gene flow species. *Journal of Heredity*, **89**, 438-450.
- Wilson EO (1959) Adaptive shifts and dispersal in a tropical ant fauna. *Evolution*, **13**, 122-144.
- Wilson EO (1961) The nature of the taxon cycle in the Melanesian ant fauna. *American Naturalist*, **95**, 169-193.
- 580 Yam RSW, Dudgeon D (2005) Inter- and intraspecific differences in life history and growth of *Caridina* spp. (Decapoda: Atyidae) in Hong Kong streams. *Freshwater Biology*, **50**, 2114-2128.

Table 1. Sample and fragment information and measures of molecular diversity.

<b>Species</b>	<b>N</b>	<b>n</b>	<b>bp</b>	<b><i>h</i></b>	<b><math>\pi</math></b>	<b><i>k</i></b>	<b># hap</b>	<b>Genbank accession numbers</b>
<i>A. lanipes</i>	185	8	773	0.352 ± 0.003	0.0005 ± 0.0005	0.421 ± 0.385	32	EU005053 - EU005083
<i>A. scabra</i>	225	9	596	0.986 ± 0.004	0.013 ± 0.007	7.714 ± 3.608	129	EU005084 - EU005224
<i>A. innocous</i>	48	4	777	0.964 ± 0.012	0.010 ± 0.005	7.850 ± 3.717	31	EU005036 - EU005052
<i>X. elongata</i>	66	5	768	0.997 ± 0.003	0.011 ± 0.006	8.109 ± 3.811	59	EU004940 - EU005000
<i>M. faustinum</i>	71	7	708	0.731 ± 0.060	0.002 ± 0.001	1.356 ± 0.853	34	EU005001 - EU005035

585 N = total number of individuals, n = number of sample sites (rivers), bp = number of COI bases used in analyses; *h* = haplotype diversity;  $\pi$  = nucleotide diversity; *k* = average number of pairwise differences; # hap = number of unique haplotypes.

590 Table 2. Population demographic parameters. *P* values are in parenthesis and significant values are in bold.

<b>Species</b>	<b>D</b>	<b>F<sub>s</sub></b>	<b>R<sub>2</sub></b>	<b>τ</b>	<b>θ<sub>0</sub></b>	<b>θ<sub>1</sub></b>
<i>A. lanipes</i>	<b>-2.683 (&lt;0.001)</b>	<b>-65.712 (&lt;0.001)</b>	<b>0.012 (&lt;0.001)</b>	0.714	0.000	0.771
<i>A. scabra</i>	<b>-2.005 (&lt;0.001)</b>	<b>-201.469 (&lt;0.001)</b>	<b>0.032 (0.003)</b>	2.130	7.059	4,026.250
<i>A. innocous</i>	-0.811 (0.222)	<b>-11.990 (0.005)</b>	0.081 (0.227)	8.936	0.000	87.930
<i>X. elongata</i>	<b>-2.139 (0.003)</b>	<b>-73.964 (&lt;0.001)</b>	<b>0.036 (&lt;0.001)</b>	8.502	0.003	175.938
<i>M. faustinum</i>	<b>-2.487 (&lt;0.001)</b>	<b>-45.119 (&lt;0.001)</b>	<b>0.021 (&lt;0.001)</b>	1.571	0.000	7.245

D = Tajima's (1989) D; F<sub>s</sub> = Fu's (1997) F<sub>s</sub>; R<sub>2</sub> – Ramos-Onsins & Rozas' (2002) R<sub>2</sub>; τ = tau, which is an index of time since the expansion expressed in units of mutational time; θ<sub>0</sub> & θ<sub>1</sub> = pre- and post- expansion values for the mutation parameter (i.e. 2Nμ, where N is the effective female population size and μ is the mutation rate per gene per generation).

595

Table 3. Inferred evolutionary processes from NCA<sup>1</sup>

<b>Species</b>	<b>Clade level</b>	<b>Inference Chain</b>	<b>Inferred evolutionary process</b>
<i>A. lanipes</i>	Total cladogram	1-2-11-12: No	Contiguous range expansion
<i>A. scabra</i>	Clade 2-13	1-2-11-17-4: No	Restricted gene flow with IBD
	Clade 3-1	1-2-3-4: No	Restricted gene flow with IBD
	Clade 3-3	1-2-3-5-6: too few clades	Range expansion or restricted gene flow
	Clade 3-8	1-2-3-4: No	Restricted gene flow with IBD
<i>X. elongata</i>	Clade 4-2	1-2-3-4: No	Restricted gene flow with IBD
<i>M. faustinum</i>	Clade 1-9	1-2-11-17:No	Inconclusive
	Clade 3-1	1-2-11-12: No	Contiguous range expansion
	Total cladogram	1-2-11-12: No	Contiguous range expansion

<sup>1</sup>NCA revealed no significant associations between geography and genealogy in *A. innocuous*.

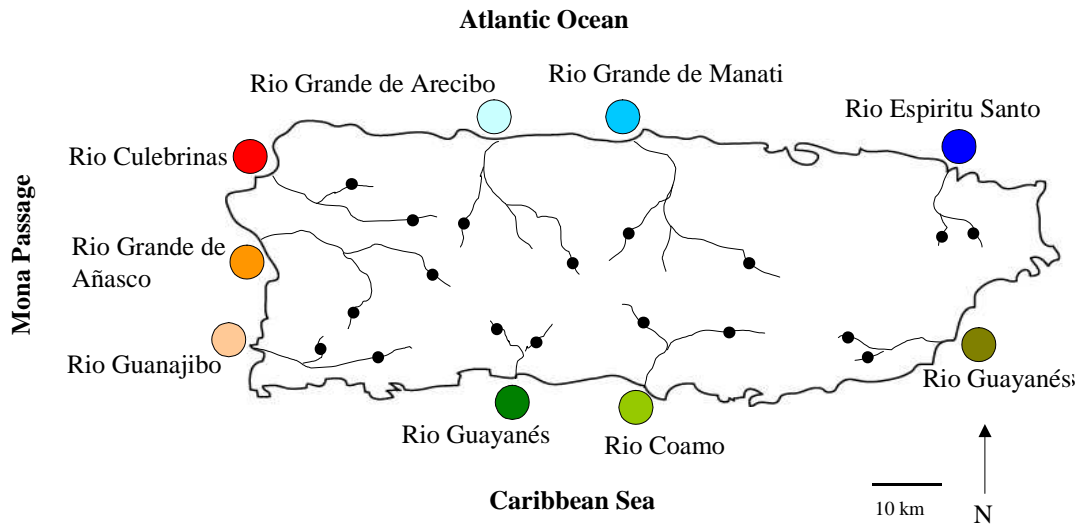


Figure 1. Map of Puerto Rico showing the sites (black dots), rivers and regions sampled. Each river is colour coded and these colours correspond to those used in the haplotype networks (Fig. 2). Similarly, each region is colour coded: Atlantic Ocean – blues; Mona Passage – reds and oranges; Caribbean Sea – greens.  
605

Figure 2. Estimates for time (in years before present) since last population expansion in each taxon at the 0.05 confidence level, showing estimated year of expansion plus lower and upper bounds. The dotted line demarks the transition from the Pleistocene (above line) to the Holocene (below line).  
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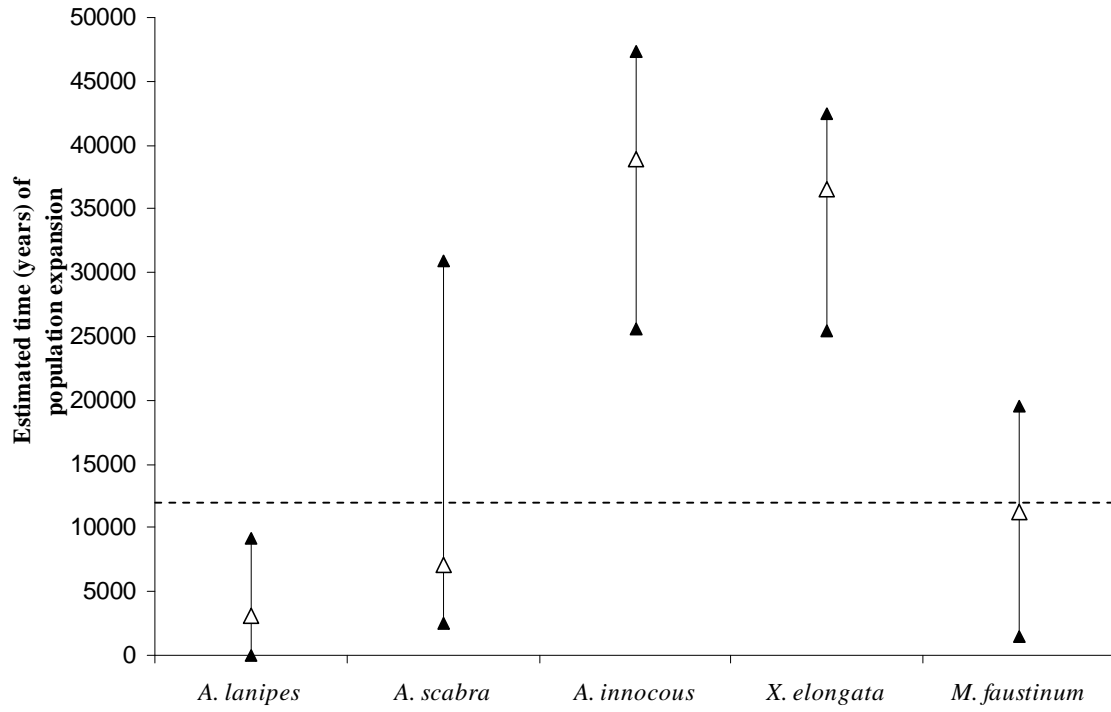
Figure 3. Haplotype networks showing nesting levels used for NCA for a) *A. lanipes*, b) *A. scabra*, c) *A. innocuous*, d) *X. elongata*, and e) *M. faustinum*. 2-step and higher clades only are shown for b-e.  
615

Fig. 1



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Fig. 2



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Fig. 3

